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PHYLOGEOGRAPHY AND HABITAT ASSOCIATIONS OF FRANKLIN'S GULLS IN THEIR UNITED STATES BREEDING RANGE

by

AnnMarie Krmpotich Bachelor of Science, Colorado State University, 2008

A Thesis

Submitted to the Graduate Faculty

of the

University of North Dakota

in partial fulfillment of the requirements

for the degree of

Master of Science

Grand Forks, North Dakota May 2012

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This thesis, submitted by AnnMarie Krmpotich in partial fulfillment of the requirements for the Degree of Master of Science from the University of North Dakota, has been read by the Faculty Advisory Committee under whom the work has been done and is hereby approved.

Dr. Rebecca Simmons

Dr. Brett Goodwin

Dr. Brad Rundquist

Dr. Marissa Ahlering

This thesis meets the standards for appearance, conforms to the style format requirements of the Graduate School of the University of North Dakota, and is hereby approved.

Dr. Wayne Swisher

<u>May 1, 2012</u> Date

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Date May 1, 2012

TABLE OF CONTENTS

LIST OF FIGURES
LIST OF TABLES ix
ACKNOWLEDGEMENTS xi
ABSTRACTxv
CHAPTER
I. INTRODUCTION TO PHYLOGEOGRAPHY AND HABITAT ASSOCIATIONS OF FRANKLIN'S GULLS1
Wetlands1
General Description1
Wetland Functions and Trends2
Wildlife Impacts
Gene Flow4
Connectivity4
Migratory Species5
Phylogeography6
The Franklin's Gull7
Population Dynamics8
The Breeding Season10
Conservation Status10

Management Needs for the Franklin's Gull	11
Study Objectives	12
Literature Cited	13
FRANKLIN'S GULL POPULATIONS REVEAL PANMIXIA DESPIT GEOGRAPHICALLY SEGREGATED BREEDING SITES	
Abstract	19
Introduction	19
Methods	24
Field Methods and Study Area	24
Laboratory Methods	25
Data Analysis	27
Results	
Haplotype Diversity and Genetic Structure	
Phylogenetic Analysis	29
Discussion	29
Literature Cited	
ASSESSMENT OF HABITAT ASSOCIATIONS OF FRANKLIN'S GULLS IN THE PRAIRIE POTHOLE REGION USING GIS AND REMOTE SENSING TECHNOLOGIES	46
Abstract	46
Introduction	47
Methods	50
Wetland Habitat Selection	50
Study Area	51
	Study Objectives Literature Cited FRANKLIN'S GULL POPULATIONS REVEAL PANMIXIA DESPIT GEOGRAPHICALLY SEGREGATED BREEDING SITES Abstract Introduction Methods Field Methods and Study Area Laboratory Methods Data Analysis Data Analysis Results Haplotype Diversity and Genetic Structure Phylogenetic Analysis Discussion Literature Cited ASSESSMENT OF HABITAT ASSOCIATIONS OF FRANKLIN'S GULLS IN THE PRAIRIE POTHOLE REGION USING GIS AND REMOTE SENSING TECHNOLOGIES Abstract Introduction Methods Wetland Habitat Selection

	Imagery Acquisition and Processing	
	Photo Interpretation	55
	Land Cover Analysis	56
Resul	lts	
Discu	ussion	
Litera	ature Cited	62
IV. CONCLUS	SIONS	71
Litera	ature Cited	74
APPENDICES		75
Appendix A.	Permits	76
Appendix B.	Digitized Maps	

LIST OF FIGURES

Figure	Page
1.	Sampling locations from the breeding range of the Franklin's gull. The letter code corresponds to sampling information (Table 2) for each colony location included in genetic analyses
2.	Phylogenetic tree produced using maximum likelihood criteria in GARLI - ln likelihood = 1701.10512. Bootstrap support is given at appropriate nodes. Each branch represents one haplotype. Parentheses prior to nodes indicate posterior probabilities. Parentheses in the resolved portion following location represent number of individuals that share the haplotype represented. The remaining 150 Franklin's gulls compose a polytomy of individuals represented among all colonies sampled. Parentheses following colony location indicate h , number of haplotypes, n , number of individuals
3.	Phylogenetic tree produced using maximum likelihood criteria in program MRBAYES $-ln$ likelihood = 1557.62088. Posterior probabilities assigned to nodes assess the strength of the relationships found at branch nodes. Each branch with resolution represents one haplotype. Parentheses in the resolved portion following location represent number of individuals that share the haplotype represented. The remaining Franklin's gulls sampled compose a polytomy of individuals represented among all colonies sampled. Parentheses following colony location indicate <i>h</i> , number of haplotypes, <i>n</i> , number of individuals44
4.	Wetland location sites with established Franklin's gull colonies $(n=4)$ paired with control sites $(n=4)$ 70

LIST OF TABLES

Table	Page
1.	Conservation status of the Franklin's gull (<i>Leucophaeus pipixcan</i>) within its U.S. breeding range by natural resource agencies, including reasons for this species' designation
2.	Sampling locations for each colony included in genetic analyses from the breeding range of the Franklin's gull. The codes in this table correspond to locations mapped in Figure 1, N = sample size, coordinates are given in decimal degrees
3.	Frequencies for 34 shared mtDNA haplotypes (115 total haplotypes) for 14 populations of Franklin's gulls sampled throughout breeding colonies found in the United States. Dashes indicated the haplotype was not found in the population
4.	Franklin's gull sampling locations (mapped in Figure 1), N(I) number of individuals, N(H) number of haplotypes, h haplotype diversity, π nucleotide diversity, and SD (standard deviation)
5.	Results of Analysis of molecular variance (AMOVA) based on Franklin's gull mitochondrial D-loop control region
6.	Description of wetland classifications measured to evaluate Franklin's gull use of wetland basins in the PPR of the U.S. in 2010; adopted from Stewart & Kantrud, (1971)
7.	Description of land cover classifications measured to evaluate Franklin's gull use of wetland basins in the PPR of the U.S. in 2010
8.	Land cover (ha) by site of four wetland basins with Franklin's Gull nesting colonies paired with four non-use wetlands
9.	Description of habitat cover of wetland basins in the PPR of the U.S measured to evaluate Franklin's Gull use in 2010

10.	Characteristics of wetland study sites: mean percentages (\pm standard deviation),
	mean road length, mean cover to water and edge to water ratios from wetland
	basins with (used) and without (unused) Franklin's gull nesting colonies across the
	PPR of the U.S. in 2010
11.	Paired t-tests results for land cover characteristics comparing wetland sites

11.	Paired t-tests results for land cover characteristics comparing wetland sites	
	with and without nesting Franklin's gull colonies during the 2010 breeding	
	season across the PPR	69

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ABSTRACT

The Franklin's Gull (*Leucophaeus pipixcan*) is a long distance migrant which is listed as a species of concern by U.S. state wildlife management agencies and in the Northern Prairie and Parkland Waterbird Plan for the U.S. and Canada. The designation of this status is due to few, isolated breeding colonies and unknown population dynamics. Few attempts have been made to address the unknown population dynamics or to quantify habitat use at breeding sites of this species. Recognizing there is a need for this information; this thesis uses deoxyribonucleic acid (DNA) evidence to show the relationships of Franklin's gull breeding colonies across the U.S. as well as quantifies habitat use through geographic information system (GIS) and remote sensing technologies. This thesis describes data collected on population structure and habitat use of Franklin's gulls. I found Franklin's gulls in the breeding range of the United States are a panmictic population. It was found habitat use of Franklin's gull colonies at breeding sites depends on multiple variables suggesting landscape management to include wetland complexes is key to persistence of nesting colonies in the United States.

CHAPTER I

INTRODUCTION TO PHYLOGEOGRAPHY AND HABITAT ASSOCIATIONS OF FRANKLIN'S GULLS

Wetlands

General Description

The dynamic nature of wetlands complicates their definition and classification. In general, wetlands are defined by plants (hydrophytes), soils (hydric soils), and frequency of flooding (Cowardin et al., 1979). A wetland is a transitional area between true aquatic habitats (e.g. rivers, lakes, estuaries, or oceans) and a dry, terrestrial (upland) habitat (Mitsch & Gosselink 2007, Dahl, 2011). A classification system for wetlands is intended to provide a standard definition at a variety of hierarchy levels to be useful to natural resource managers for the purpose of science, education, and administration. Although many wetland classification systems have been developed, the one most commonly adopted for natural resource management in the U.S. today is the "Classification of Wetlands and Deep Water Habitats" (Cowardin et al., 1979). At a regional scale for the prairie pothole region (PPR), the classification system most widely adopted is the "Classification of Natural Ponds and Lakes in the Glaciated Prairie Region" (Stewart & Kantrud, 1971). Due to its practicality, this system is favored by many field biologists. Seven major classes of wetlands are distinguished by vegetational zones which include: Class I (ephemeral ponds), Class II (temporary ponds), Class III (seasonal ponds and

lakes), Class IV (semi-permanent ponds and lakes), Class V (permanent ponds and lakes), Class VI (alkali ponds and lakes), to Class VII (fen ponds).

Wetland Functions and Trends

Wetlands serve many valuable functions on the landscape; including ecological, economic, and social benefits. Wetlands serve to mitigate floods, recharge ground water, protect shoreline, and cleanse polluted water coming from surrounding lands. These functions are lost when destruction, degradation, or any form of alteration occurs; additionally negatively impacting biotic factors (Mitsch & Gosselink, 2007). Public recognition of the value of wetlands has triggered a movement of protection over the past 30 years. Despite progress towards wetland protection, there still exists continuing pressure from both anthropogenic and environmental stressors that threaten these habitats. At the time of European settlement, there were an estimated 221 million acres of wetlands in the conterminous U.S. (Dahl, 1990, Dahl, 2000). From the 1950s to the 1970s the average annual loss of wetlands was 380,000 acres (Frayer et al., 1983). With the enactment of The Clean Water Act (1972) and the Food Security Act of 1985 ("Swampbuster"; P.L. 99-198), wetland drainage slowed considerably (Reynolds et al., 2006). However, wetlands are still being lost at alarming rates; the amount of wetland habitat loss annually in the U.S. is 58,500 acres (Dahl, 2000). As of 2009, it was reported the total wetland acreage was 110.1 million (Dahl, 2011), which is less than half of the pre-settlement wetlands found in the continental U.S. With the continued loss of freshwater wetlands come severe consequences in hydrologic and ecosystem connectivity (Dahl, 2011).

2

Wildlife Impacts

Wetland habitats have been lost or degraded due primarily to agricultural practices. There is a conflict of interest for wildlife managers between locations of land optimal for supporting wetland wildlife and land with high agricultural value being drained and converted to crops. Severely affected by these wetland alterations are the wildlife species in the PPR. One major concern in wildlife management is breeding bird populations, as the PPR is a geographical 'hot spot' for migratory bird use during the breeding season, and available habitat will ultimately determine their success. Loss of wetland habitats has negative consequences to wetland bird populations (Johnson, 2001). In particular, wetlands play a critical role in migratory bird breeding habitat and the loss of this habitat has severe impacts to these populations. Consequently, populations of many waterbird species have been designated as a species of concern (Beyersbergen et al., 2004).

Understanding the impacts of landscape change is important, especially for migratory species that rely on a mosaic of habitats to meet their annual needs during the breeding, non-breeding, and migration periods (Skagen et al., 1999; Drake et al., 2001). Reductions in the availability of wetland stopover sites could negatively impact migratory bird populations by reducing the overall quality of the landscape and increasing the distance between suitable stopover sites (Skagen, 2006, Smith, 2008). There is much concern for migratory bird species that rely on wetlands in the PPR either as stopover sites or for their breeding habitat as the availability of quality wetlands within this landscape continues to decline.

Gene Flow

Connectivity

Habitat loss through anthropogenic landscape alteration (conversion of wild lands to logging, development, or agricultural practices) can result in isolated habitat patches which results in a significant impact at the population level for wildlife species (Funk et al., 2010). Habitat loss can segregate subpopulations even further, leaving them vulnerable to reduced genetic diversity (Howes et al., 2009). Segregated populations often have reduced numbers of individuals, increasing the chance of genetic drift (Evans et al., 2008), reduced genetic variation within a population, and increased genetic diversity between populations, possibly leading to speciation (Irwin et al., 2001). The consequence of this segregation and reduction in population size is a loss of local alleles (Lacy, 1987). Importantly, changes in gene frequency can occur over a relatively short period of time (Rolshausen et al., 2009), which has implications for population survival and conservation. Some concerns of gene frequency change occurring quickly is a bottleneck, inbreeding, and the population's ability to evolve in response to environmental change- all directly affects fitness. In light of these considerations, understanding how populations within a species are distributed across its geographic range is important in conservation biology and to species management.

Genetic monitoring at various spatial scales can assist in detecting changes in species abundance and diversity. Situations where this is important include species designated as a conservation priority and threatened and endangered (T & E) species where this information can help determine both management strategies and intensity. Further, determining the amount of connectivity among sub-populations can allow inferences about the status of population inter-mixing from panmixia to geographic isolation, which impacts management approaches. Managing disjunct populations of species is crucial, as survival at a regional scale often depends on population growth and dispersal characteristics at the local scale (Fahrig & Merriam, 1994). Successful management of disjunct populations includes fine-scale determination of what types of areas and protections are necessary; whereas, management of a panmictic species where there are no restrictions in the population for mating (i.e., all individuals are potential mates), will involve different strategies including conservation prioritization of a species' needs at various life stages, preserving intraspecific genetic diversity, and protecting the ecological and evolutionary processes necessary for the species' persistence.

Migratory Species

Understanding genetic diversity is key to determining metapopulation dynamics and for developing species conservation plans for migratory species (Esler, 2000; Taylor, & Norris, 2010). A migratory species' range depends on habitat availability, including necessary stopover areas (Skagen, 2006) and on a species' response to climate in both wintering and breeding grounds (Mustin et al., 2007). Migratory birds depend on stopover sites to refuel their energy stores depleted during travel between distant summering and wintering areas (Weber, 1999). If major stopover points are lost (not available), migratory behavior must adapt in order to maximize fitness (Smith & Deppe, 2008; Weber, 1999). New migration routes are known to occur with changing environmental conditions (Sutherland, 1998). However, the extent to which species are capable of shifting migration patterns to occupy new ranges is a major concern, especially given predicted impacts of global climate change (Mustin et al., 2007). Most migratory species, although dispersed over a large geographic area, exhibit some degree of genetic population structure (Avise & Hamrick, 1996; Webster et al., 2002). Causes of population structure are the result of restricted gene flow, due to physical or behavioral barriers that act to isolate breeding among subpopulations (Slatkin, 1987). The movement of individuals and their genes influences a number of ecological processes including population persistence and adaptive response to environmental change (Frankham et al., 2002). The use of genetic approaches in wildlife management can help address complex species management issues, such as deciphering meta-population dynamics, provided a basic knowledge of gene flow is known.

Phylogeography

The study of a species' geographic distribution and the study of a species' evolutionary patterns are known as phylogeography (Avise, 2000). Phylogenetic approaches can provide a myriad of ecological applications, from hypothesis testing to conservation planning. Previous studies have used phylogenetic methods to decipher a species range (Zeisset & Beebee, 2001), identify biogeographic barriers (Sonsthagen et al., 2011), distinguish movement patterns of migratory species (Mehl et al., 2004), provide insight to landscape genetics (Oomen et al., 2011), and assist in endangered species management (Lei et al., 2003). The use of phylogeography can provide important information for conservation purposes through determining population structure and by providing evidence for demographic events, such as changes in population size or dispersal (Avise, 2000).

The Franklin's Gull

The Franklin's gull (*Luecophaeus pipixcan*) is a species that utilizes wetlands in the PPR and exhibits segregated populations, an ideal candidate to investigate habitat loss in a migratory species for this region. The Franklin's gull relies on Class IV & V wetlands (Stewart & Kantrud, 1971) for nesting which are the predominant wetland types, in terms of total acreage, throughout the PPR. The wetland nesting locations in this region used by Franklin's gulls are geographically segregated. The Franklin's gull is a migratory species, traveling up to 10,000 km between its breeding areas in North America to wintering areas along the coasts of central Peru and northern Chile (Burger & Gochfeld, 2009). It is only one of two species of North American gulls to migrate south of the equator (Burger & Gochfeld, 1994).

The Franklin's gull arrives on its breeding grounds in mid-to late April and remains until August (Soos, 2004, Burger & Gochfeld, 2009). Following the nesting period, in late July the species makes broad, multi-directional post-breeding movements throughout its breeding range. Between September and early October, larger flocks begin to assemble as they prepare for their departure to southern wintering areas (Burger & Gochfeld, 2009). The Franklin's gull winters primarily along the western portion of coastal South America with a small isolated inland wintering population in Peru (Burger, 1996; Burger & Gochfeld, 2009). Due to the dynamic nature of prairie wetlands and their vulnerability to both periodic drought and drainage, breeding colonies of Franklin's gulls often shift colony locations from year to year, and in some years local populations are suspected to forego annual breeding opportunities altogether (Burger and Gochfeld, 2009). Franklin's gull breeding populations on wetlands have become reduced and segregated over time, likely because of habitat loss and increased human disturbance at the breeding sites (Burger & Gochfeld; 1994). During the Dust Bowl era (1930s) most of the historic breeding sites were lost as a result of large-scale drainage projects but now with the establishment of state and federally protected wetlands (national wildlife refuge system; NWR) Franklin's gulls nest almost exclusively on public lands (Burger & Gochfeld, 1994).

Population Dynamics

The restoration and creation of large wetland complexes, mainly on protected national wildlife refuges over the past century has aided in expansion of Franklin's gull populations (Burger & Gochfeld, 1994), however population size and dynamics remain uncertain and a conservation concern. Since the early 1900s, the largest Franklin's gull breeding colonies in the U. S. (colonies of more than 100,000 breeding pairs) have been located on marshes in the prairie pothole region (PPR), primarily in North Dakota and Minnesota (Burger & Gochfeld, 1994). Segregated small sub-populations are located in Oregon, Nevada, and Utah (Burger & Gochfeld, 1994). In any given year there are less than 50 colony site locations across the North American breeding grounds for the continental population (Burger & Gochfeld 1994) with approximately 15-20 sites of those located in the U.S.

A paucity of information exists on total population size at both the continentallevel and within the Prairie & Parkland Region for some colonial nesting waterbird species, according to *The Northern Prairie and Parkland Waterbird Conservation Plan* (Beyersbergen, 2004). Of those colonial species, the Franklin's gull has been estimated as approximately 2.5 million birds in the Northern Prairie & Parkland Region during migration (Beyersburgen, 2004). Another estimate of the global population of Franklin's gull was 315,000-991,000 adults (Milko et al., 2003). However, exact numbers are unknown, due to this species' typically remote nesting locations and an inability to assess population status from standard surveying methods (e.g. North American Breeding Bird Survey).

Detection of these colonies relies heavily on visual identification of colonies or observation of concentrated numbers of gulls in foraging areas (e.g. tilled farm fields in spring), indicating a possible colony in the vicinity. The most common and widespread survey methods to monitor the status and trends of North American bird populations (e.g. North American Breeding Bird Survey), are typically conducted roadside. A roadside survey methodology is often of limited value for detection of colonial nesting waterbird species that utilize marsh habitats. Colonies of these species, including those of Franklin's gulls, are typically located on the interior of wetlands within dense vegetation and away from areas of potential disturbance (e.g. roadsides, uplands, Burger & Gochfeld, 2009). Franklin's gull behavior exacerbates these issues. Adults rarely wander away from a nesting colony when eggs and/or chicks are present (Burger, 1974), which decreases the chance of detecting colonies. Despite these sampling limitations, some literature suggests that continental populations of Franklin's gulls have declined. Breeding Bird Survey data from 1968-1991 suggests an overall 90% decline (Sauer et al., 2008) but causes of this decline are unknown and trends in recent surveys show conflicting results (Beyersbergen, 2004; Kushlan, 2004).

The Breeding Season

The Franklin's gull requires large semi-permanent and permanent freshwater marshes with emergent vegetation and open water (Beyersbergen, 2004) on its breeding range. They are a colonial nesting species that build nest structures in the form of over water platforms or utilize existing muskrat houses (Burger & Gochfeld, 1994). These nests require continual maintenance by adding vegetation to prevent sinking or flooding until young-of-the-year gulls reach fledging stage. Nest characteristics vary between colony sites: types of vegetation (*Typha* spp., *Schoenoplectus* spp.); density of vegetation (live and dead); and location of nests within each wetland (center versus edge; Burger, 1974). Quantified land cover habitat characteristics which serve a role in nest site selection at breeding colony localities of Franklin's are currently limited for this species.

Clutches range from two to four eggs (mean n=3) (Burger & Gochfeld, 2009); only one brood is reared each season (Burger & Gochfeld 2009). Duration of pair bonds is unknown (Burger & Gochfeld, 2009). Parents remain in close vicinity of nests (within 30km) when searching for food (Burger & Gochfeld, 2009; Beyersbergen, 2004). It is common to see Franklin's gulls in flocks when searching for food, usually over water or in agriculture fields; dominating their diet composition is earthworms, grubs, grasshoppers and midges (Chironomidae; Beyersbergen, 2004).

Conservation Status

Nesting colonies throughout the breeding range of Franklin's gulls share a common threat: habitat loss and disjunct populations. Across North America, breeding populations of Franklin's gulls are threatened by habitat loss and other human encroachments on wetlands (Hagen et al., 2005; Bakker, 2005; MNDNR, 2008; Idaho

Department of Fish and Game, 2005; Montana Fish, Wildlife, and Parks, 2009; Ivey & Herziger, 2006). The Franklin's gull was recently listed as a species of high concern in the Northern Prairie and Parkland Region (Beyersbergen, 2004). Within its breeding range in the U.S., the Franklin's gull is currently listed as a species of conservation concern in all states where nesting colonies occur. Table 1 summarizes the status of Franklin's gull throughout its U. S. breeding range by state.

Management Needs for the Franklin's Gull

Management of Franklin's gulls is hampered by a lack of information about habitat preferences and population structure. Colonial waterbird species are dependent on relatively few critical wetland sites within their breeding range. With naturally fluctuating water levels of prairie wetlands due to periods of drought and deluge, individual sites may not serve as suitable nesting sites in consecutive years. The ephemeral nature of these sites forces colonial nesting waterbird species to relocate to other nearby locations or to forgo reproduction during a given year. This complicates resource managers' efforts to analyze habitat use at existing nesting sites and to determine criteria for wetland preservation at nesting areas. Quantification of habitat use at successful nesting colony locations may assist in these efforts. Further, there are no data on population structure for the Franklin's gull; understanding the relationships among colonies within their breeding range provides critical information about population dynamics. As Franklin's gull populations may be forced to relocate to different breeding areas annually, it is particularly important to understand the relationships of individuals to one another.

Study Objectives

Population genetics is a powerful tool for wildlife managers when coupled with spatial data (Hitt et al., 2003; Haig et al., 2004; Schwartz et al., 2007; Kendall et al., 2009). Understanding how a species is connected across its geographic range is important, especially given ongoing continental declines in optimal wildlife habitat. Further, spatial information can identify key habitat elements for successful breeding grounds for a species. The goal of this study is to investigate the relationship among landscape composition at breeding sites and population genetic structure of the Franklin's gull in the U.S. To achieve this overall goal, the objectives of this study are to: 1) estimate the genetic diversity among and between Franklin's gull colonies on U.S. breeding grounds to infer levels of philopatry and dispersal, and 2) quantify habitat land cover at nesting sites to determine use trends at successful nesting colony locations. A combination of findings related to both objectives will provide information for landscape management approaches for Franklin's gull breeding colonies within the U.S.

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State	Conservation Status	Implications	Reference
ND	Level 1: Species in greatest need of conservation	The state has one of the largest breeding colonies in the world	Hagen et al. 2005
SD	Level 1: Priority bird species	High maximum abundance of the species within its range in South Dakota; The species is showing population declines in the state/across its range	Bakker 2005
MN	Special concern	Extremely uncommon in the state and has unique or highly specific habitat requirements that deserve careful monitoring of its status	Minnesota Department of Natural Resources 2008
MT	Species of concern	Native taxa at risk to state extirpation due to declining population trends, threats to their habitats, and restricted distribution	Montana Natural Heritage and Montana Fish Wildlife and Parks 2009
ID	Protected nongame; Imperiled breeding	Habitat threats and disjunct populations; habitat stresses including fluctuating water levels, exotics.	Idaho Department of Fish and Game 2005
UT	Priority species of moderate concern	Breeding, migrant of the state. Fluctuating, uncertain breeding trend, need for population assessment. Management issues primarily concerned with providing ideal colony site conditions.	Ivey and Herziger 2006

Table 1. Conservation status of the Franklin's gull (*Leucophaeus pipixcan*) within its U.S. breeding range by natural resource agencies, including reasons for this species' designation.

CHAPTER II

FRANKLIN'S GULL POPULATIONS REVEAL PANMIXIA DESPITE GEOGRAPHICALLY SEGREGATED BREEDING SITES

Abstract

Managing species with segregated populations is a challenging task and requires an understanding of population structure. For segregated populations, the conservation of genetic diversity is especially important. As habitat fragmentation increases, so does the importance of genetic diversity. Within this chapter I assessed genetic diversity among and between segregated colonies of Franklin's gull to understand population structure across their breeding range in the U.S. To determine genetic variation and gene flow among these sub-populations, approximately 330 base pairs of the mitochondrial DNA (mtDNA) D-loop were amplified and sequenced. Samples represented 208 individual Franklin's gulls from 14 separate breeding colonies. I used likelihood and Bayesian algorithms to examine phylogeography and population structure, that revealed high levels of genetic diversity among segregated populations. I found complete lack of population structure for this widely dispersed waterbird, indicating low levels of natal site fidelity. I concluded that Franklin's gulls be managed as one panmictic population across the U.S. rather than distinct sub-populations.

Introduction

Many migratory species are dispersed over a large geographic area during their breeding and wintering periods, yet exhibit some degree of population structure (Avise &

Hamrick, 1996; Webster et al., 2002). The movement of individuals and their genes influences a number of ecological processes including population persistence and adaptive response to environmental change (Frankham et al., 2002). Causes of restricted gene flow that lead to population structuring can be due to physical or behavioral barriers that act to isolate breeding potential among subpopulations. Habitat loss can segregate subpopulations even further, leaving them vulnerable to reduced genetic diversity (Howes et al., 2009).

Understanding how a species is geographically distributed can help provide information at various scales for management purposes. Managing disjunct populations is important because survival at a larger, regional scale often depends on growth and dispersal characteristics at the local scale (Fahrig & Merriam, 1994). Management concern and decisions will be different for a species exhibiting separate populations (i.e., sub-populations) compared to population panmixia. Management concerns of species exhibiting panmixia include reduced genetic diversity over time, how a species responds to fluctuating environmental conditions, lower adaptive potential, and the opportunity for a significant portion of a population to be impacted by disease (Avise, 2004). Panmixia can offer a greater potential of mate choice and reproduction with the high level of interconnectedness among individuals within a population. These factors directly impact a species at the local level which ultimately impacts the global population.

For migratory species, a high degree of female philopatry to natal sites is necessary for mtDNA to be conserved (Mortiz et al., 1987; Miller-Butterworth et al., 2003; Avise, 2004). Thus, to maintain genetic diversity in the case of high natal philopatry, segregated subpopulations of a migratory species are recognized and sometimes managed as separate units, since intermixing among subpopulations is unlikely to occur.

Management concerns for maintaining genetic diversity for migratory colonial nesting waterbird species is especially important given they have large, dense nesting colonies but depend on a select few breeding localities. Semi-permanent and permanent wetland quality and availability should be a management focus, as these species are more likely to exhibit a higher degree of site fidelity (Prevot-Julliard et al., 1998; Kushlan et al., 2002). Colonial birds that concentrate in large numbers of select wetlands experience increased threats and vulnerability to more individuals compared to species that are non-colonial. For example, the occurrence of predation, encroachment of invasive species, destruction or degradation of a wetland (e.g. sedimentation decreases invertebrate diversity), pollutants, or disturbances (human use too close to nest sites: boats, vehicles, agriculture machinery) have potential negative impacts that result in an exponentially larger effect than would be predicted with solitary nesting sites (Kushlan et al., 2002). These variables emphasize the importance of understanding population structure for informing management practices of colonial nesting migrants.

The value of agricultural land within the PPR has resulted in conversion of more than half of the historic eight million hectares of wetlands (Dahl & Johnson, 1991). Between 2004 and 2009, the total wetland loss was estimated to be 25,210 hectares (a 140% increase of wetland loss over 1998-2004) bringing the nation's total wetland acreage to just over 44.5 million hectares (Dahl, 2011). Remaining wetlands in the PPR

21

continue to face threats, including intensification and expansion of agricultural production, development, and climate change (Johnson et al., 2005).

As a result of wetland loss, most of the existing wetlands that are used by nesting colonies of Franklin's gulls are managed impoundments with associated water control structures, dikes, and ditches. Water levels are managed and monitored to maintain a variety of wetland habitats; primarily in the PPR they are to meet waterfowl production objectives (Fredrickson & Reid, 1988). Management of these impoundments also often include regularly scheduled draw-downs. During this management technique water from wetlands is drained periodically to improve seed production, increase invertebrate abundance, and perform habitat maintenance such as disking, grazing, or burning (Cross & Vohs, 1988). Habitat requirements and more specifically water level preferences, differ among waterbird species (Kaminski et al., 2006). Often times many species of different guilds utilize the same wetland basin. Optimal depths for foraging of shorebird species such as sandpipers (*Calidris* sp.) and yellowlegs (*Tringa* sp.) are less than 8 cm whereas most dabbling duck species range from 8-23 cm. Many species need deeper water for nesting purposes, as is the preference of redheads (Aythya americana) at 61 cm of water (Fredrickson, 1991). Thus, managed wetlands often have different water levels depending on the time of year trying to meet the needs of a variety of phenological events but not all species preferences can be met.

All of these factors are important when considering the status of the Franklin's gull. The Franklin's gull is a migratory species, traveling up to 10,000 km between its breeding areas in North America to wintering areas along the coast of central Peru to northern Chile. It is only one of two species of North American gull to migrate south of

the equator. The Franklin's gull arrives on the breeding grounds in mid to late April and remains until August (Soos, 2004; Burger & Gochfeld, 2009). Following the nesting period, in July the species scatters and wanders throughout the breeding range in all directions. By September to early October larger flocks begin to assemble as they prepare for their departure south (Burger & Gochfeld, 2009).

The species breeds colonially on large, deep water marshes (0.3-0.6 m) with an interspersion of emergent vegetation for construction and attachment of floating nests. Mis-timed management of water levels can lead to failure to establish or abandonment of Franklin's gull colonies (Burger & Gochfeld, 2009). In the Northern Prairie and Parkland Region the Franklin's gull is listed as a waterbird species of High Concern, due to inadequate population information and large portions of the continental population using this region (Beyersbergen et al., 2004). Population trends for Franklin's gull are not well understood, partly because the remote breeding locations make population estimates logistically difficult. Understanding the genetic population structure of the Franklin's gull is important because it will aid in the overall understanding of the structure of colonial breeders and serve to inform management decisions, especially given population concerns, disjunct populations, and threats of further habitat loss.

In this study I used mitochondrial DNA (mtDNA) markers to gain insight into Franklin's gull population structure. MtDNA is a valuable tool for examining population structure. Successful uses include the elucidation of butterfly migration patterns (Salazar et al., 2008), range expansion in fish (Coscia & Mariani, 2011) and bird philopatry (Avise et al., 1992). The mitochondrial region D-loop was chosen based on its rapid rate of known mutation compared to other markers. MtDNA lacks recombination found in nuclear DNA and is uni-parentally inherited (Avise, 2004) making it useful to track recent population relationships.

My objectives were to document the genetic differentiation within and between active Franklin's gull breeding colonies throughout the U.S. by using the mitochondrial D-loop to estimate levels of gene flow. Based on other species, I predicted that Franklin's gull colonies would exhibit population structure across the range sampled. To examine this observation, I determined the level of dispersal and phylogeographic population structure of Franklin's gulls breeding in the U.S. In addition I determined the level of genetic diversity within and between each individual nesting colony to inform more effective management practices. I also predicted to find population structure among subpopulations to the east and west of the Rocky Mountains, due to philopatry and based on previous work with other avian populations (Burg et al., 2003).

Methods

Field Methods and Study Area

During the 2012 breeding season, Franklin's gull feather samples (molted feathers on floating nest platforms) were collected from fourteen nesting colonies across the species' U.S. breeding range (Fig. 1). To collect a representative sample of Franklin's gulls found at each colony, nests were chosen randomly throughout the colony. Once feathers (1-5) were collected from a nest bowl, they were placed in uniquely labeled paper coin envelopes for storage until processing. All feathers were stored at room temperature. Proper permits were obtained and a proper feather collection and possession protocol was followed (Appendix A). The sampling sites for feather collection included: Thief Lake Wildlife

Management Area (WMA), MN; J. Clark Salyer National Wildlife Refuge (NWR), ND; Lake Alice NWR, ND; Beaver Lake Waterfowl Production Area (WPA), ND; Rush Lake WPA, ND; Sand Lake NWR, SD; North Detroit Township, Brown County, SD; Benton Lake NWR, MT; Bowdoin NWR, MT; Knutson's Bay, MT; Red Rock Lakes NWR, MT; Grey's Lake NWR, ID; Oxford Slough WPA, ID; and Bear River Migratory Bird Refuge, UT (Fig. 1, Table 2). The number of nests feathers were collected from was approximately 50 at each colony location. This number was calculated based on the number of viable sequences needed and buffered against laboratory processing errors. The number of nests which were processed to represent individual DNA sequences was determined using similar studies investigating avian population structuring through DNA evidence (Oomen et al., 2011; Reudink et al., 2011; Eo et al., 2010; Barrowclough et al., 2004); in this literature a range of 4-35 individuals per sampling location was used. Sample size by site in this study ranged from 6-20 (Table 2).

Laboratory Methods

The calamus (quill) was clipped off from the rest of the feather shaft for collected feather samples. MtDNA was extracted from the blood in the calamus using a Qiagen DNeasy kit using standard protocols (QIAGEN Inc., Valencia CA, USA). Each extraction set had a blank control that did not contain feather material to test for contamination across samples. Following successful DNA extraction the DNA was suspended in EB buffer and stored in a -70°F freezer. DNA was quantified in the extraction samples using a Nanodrop Spectrophotometer (ND 1000). A portion of the D-loop of mitochondrial DNA (approximately 330 base pairs) was selected for this study. The mitochondrial region D-loop was chosen based on its rapid rate of known mutation compared to other markers. Further, mtDNA lacks recombination found in nuclear DNA and is uni-parentally inherited (Avise, 2004). Because of these factors, mtDNA is able to track recent population relationships. Two primers were manually designed for this study using existing sequences deposited in NCBI Genbank, accession numbers: FM209692-FM209694 (Sternkopf et al., unpublished) using conserved regions, specifically HVR-1 region of the D-loop; Betty (forward: GGAGGTTTACATTAACCTAT) and Fred (reverse:

CTAGCTTCAAGACCATAC).

Polymerase chain reaction (PCR) reactions were performed with the Ex Taq Kit (Takara biotechnology Co., Ltd) using standard procedures in an Eppendorf thermocycler (Eppendorf, Hamburg, Germany). The cycling conditions for the PCR reaction was 1 minute at 94°C (DNA denaturation), 1 minute at 47°C (DNA annealing), and 1 minute at 72°C (DNA elongation). This process was repeated for 29 cycles. One final cycle of 1 minute at 94°C, one minute at 47°C, and 10 minutes at 72°C (final elongation phase) completed this process (Simmons & Scheffer, 2004). Samples were maintained at 4°C until further processing.

PCR products were visualized using gel electrophoresis on a 2% agarose gel containing 0.1 µg/ml ethidium bromide via UVP Bio-Imaging System (Cambridge, UK) under an AutoChemi ultraviolet transilluminator. Successful amplifications were cleaned using a Qiagen purification kit (Santa Clarita, CA) according to the manufacturer's standard protocols. Purified PCR products were sequenced using a Big Dye Xterminator

kit and visualized with ABI 3100 capillary sequencer. Sequences have been deposited in NCBI Genbank (accession numbers: to be submitted).

Data Analysis

Sequence data were verified via NCBI GenBank using a BLAST search to ensure no contamination from other possible sources of DNA. Sequences for individuals were assembled with Sequencer 4.6 (GeneCodes Corp.) and consensus sequences for each sampled individual were manually aligned by eye. Sequences for individual Franklin's gulls were assembled to produce a consensus sequence for an individual bird assuming feathers found in a nest bowl were from the same individual. Consensus sequences for each individual were aligned and manually adjusted when appropriate. Mutations at each position were verified with original chromatograms for each individual. All individual Franklin's gull sequences were complied into a matrix for phylogenetic analyses. Haplotypes (individual DNA sequences) were identified using the statistical parsimony approach in program TCS version 1.21 (Clement et al., 2000).

Haplotype frequency data were analyzed in program ARLEQUIN v. 3.5.1.2 (Excoffier, 2010) to obtain F_{ST} (gene flow) and nucleotide diversity (π) to infer the level of gene flow among and between populations. Arlequin was also used to perform exact tests of population differentiation (Raymond & Rousset, 1995), as well as haplotype diversity to measure the uniqueness of a particular haplotype within each population and analysis of molecular variance (Excoffier et al., 1992).

Two related methods were used to select the appropriate model for phylogenetic analyses. For likelihood analyses, model selection was obtained using Akaike Information Criterion (AIC) as the criterion in Modeltest 3.7 (Posada & Crandall, 1998; Posada & Buckley, 2004). Model selection was based on Bayesian Information Criterion (BIC) scores and performed using Mr.Modeltest v. 2.2 (Nylander, 2004) for Bayesian searches and executed in program PAUP (Posada & Crandall, 1998). The most appropriate model for both algorithms was GTR+I+G.

Phylogenetic trees were generated using maximum likelihood criteria in GARLI (Genetic Algorithm for Rapid Likelihood Inference; Zwickl, 2006) and MRBAYES (Huelsenbeck & Ronquist, 2001; Ronquist & Huelsenbeck, 2003). Maximum likelihood was used to search all possible combinations of tree topology and branch length. Maximum likelihood trees were constructed in GARLI using 1,000,000 generations of trees after likelihood scores became stationary. Resulting nodes of the final tree were assessed by re-sampling the matrix for 1,000 parametric bootstrap replications. Bayesian methods were used to produce a phylogenetic tree assuming posterior probability distribution. Eleven million generations of random trees were generated in MRBAYES, using two searches with four Markov Chain Monte Carlo chains (one heated, three cold). Cold chains were sampled every 100 runs. Analyses were determined when less than 0.01 difference between split frequencies of the two searches were reached. Twenty-five percent of resulting trees were discarded as burnin, as recommended (Huelsenbeck & Ronquist 2001; Ronquist and Huelsenbeck 2003). Posterior probabilities were used to assess the strength of relationships.

Results

Haplotype Diversity and Genetic Structure

Of the 208 individual Franklin's gull mtDNA D-loop sequences sampled, 115 unique haplotypes were present. Thirty-four haplotypes had more than one individual assigned to it (Table 3), whereas the remaining haplotypes were only found in one sample. The amount of unique sequences indicates a high degree of genetic diversity. Each of the 14 sub-populations (colonies) contained a variety of haplotypes ranging from 6-17 (Table 4). Arlequin revealed high haplotype diversity (h = 93.4-100%) across all sub-populations and high nucleotide diversity (0.0-4.2%) measuring the mtDNA sequences of individual birds within each colony (Table 4). Pairwise distances between all individuals revealed high diversity, ranging from 0.0-4.2%. The AMOVA showed low variation (0.06%) among sub-populations, but high genetic variation (99.40% explained) within groups (Table 5). Pairwise F_{ST} values between colonies were very low (Table 5) as was overall F_{ST} (0.00597, p < 0.05). F_{ST} values range from 0 (population panmixia) to 1 (two separate populations). This indicates populations sampled are interbreeding freely.

Phylogenetic Analysis

The haplotype network from GARLI revealed few resolved clades (-*ln* likelihood = 1701.10512). Bootstrap support ranged from 54-60% (Fig. 2). Bayesian methods also were unable to resolve haplotype groupings (-*ln* likelihood = 1557.62088). Posterior probabilities ranged from 57-91% (Fig. 3) for supported relationships. Overall, populations did not show genetic structure based on sampling location or any spatial scale thus indicating individuals sampled represent a panmictic population.

Discussion

Contrary to our predictions, we found high levels of genetic variation (115 haplotypes in 208 individuals across 330bp) and no population structure across the U.S. breeding areas for this species (Figs. 2 and 3). These results suggest a lack of natal site fidelity for the population, as a whole. Although Franklin's gull breeding sites are geographically isolated in their breeding range, congregations of large numbers of Franklin's gulls on the wintering area have been documented (Burger, 1996), likely providing opportunities for population mixing to occur. Given the ephemeral nature of prairie wetlands, the benefits of returning to the same wetland in subsequent years are likely diminished (Covich et al., 1997); therefore, spring migration may occur in flocks of mixed breeding origin, whereby a female of one breeding location may follow those of another to different area, depending on year and previous nest and chick success (Haas, 1998; Catlin et al., 2005; Lecomte et al., 2008; Robinson & Oring, 1997).

Other waterbird species with similar colonial nesting traits display a lack of genetic structure. American white pelicans (*Pelecanus erythrorhynchos*) are long distance migratory waterbirds, nest colonially on inland sites, and are geographically segregated across their breeding range. MtDNA evidence indicates that American white pelicans are a panmictic population; reasons suggested include having low natal site fidelity, high rates of mobility, and lack of reproductive isolating barriers (Oomen et al., 2011, Reudink et al., 2011). An examination of lesser snow goose (Chen caerulescens) genetic structure found that individuals of this species, despite having disjunct populations in the breeding range, come together on the wintering grounds and intermix during the mate selection period (Cooke et al., 1975), therefore, gene flow is maintained through males returning to a female's natal site. Despite a recent population decline, Peruvian booby (Sula *variegata*) populations have shown to maintain high levels if genetic diversity and are genetically panmictic (Taylor et al., 2011) through high levels of dispersal and interaction due to ample availability of breeding locations, and opportunities for genetic exchange with many individuals for this colonial seabird.

Movement of just a few individuals between populations can be enough to prevent genetic differentiation among populations (Lacy, 1987; Oyler-Mccance, 2005). A general rule of thumb in conservation biology for long-term persistence of a species is one migrant individual per generation (OMPG rule), for isolated populations. The OMPG rule states one migrant into a subpopulation per generation is sufficient to maintain healthy levels of gene flow between isolated populations (Mills & Allendorf, 1996). Dispersal among breeding populations reintroduces genetic variation to subpopulations, causing within-subpopulation heterozygosities to stabilize after an initial rapid decline in genetic diversity. Genetic variation is required for a population to adapt to changing environments, new predators, diseases, parasites, changing climatic conditions, competitors, and changing food supplies (Lacy, 1987). Other species such as spiny lobster (*Palinurus gilchristi*, Naro et al. 2011; Tolley et al. 2005), bees (Beveridge & Simmons, 2006), and fish (White et al., 2009) reveal panmictic populations despite widely dispersed populations and long migratory paths.

Though mtDNA is informative for examining population genetic structure, additional markers (microsatellites) would prove beneficial for examining mate fidelity and annual pair formation. Further, expansion of sampling Franklin's gulls in the northern extent of their breeding range in Canada would further augment these results in determining if the trend it throughout or unique to the southern extent of the Franklin's gull breeding range. Additional studies regarding adaptive potential to a changing climate, and particularly the effects it may have on reproductive success are needed. More research on both wintering and breeding areas would facilitate a more comprehensive approach to management of this species in the breeding areas. Expanding results to encompass the Canadian breeding range, as well as wintering sites would provide information over the entire range of the species. It would also help to identify timing of pair formation and detect dispersal patterns.

Results of this work will aid in informing future management decisions. Based on the aforementioned results, Franklin's gull management decisions can be made with an understanding that each colony of breeding adults is genetically similar. The populations sampled have a high genetic diversity revealing healthy levels of population interaction during some point in this species' annual cycle. Given the concentrated use of few select wetlands, a mosaic of wetland habitats will be important for maintaining breeding localities and offer the Franklin's gull more options to establish a successful nesting colony. Identifying preference of why sites are chosen for nesting among breeding areas will provide a better understanding for this long distance migrant's habitat needs and ultimately the species' long-term success.

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Sampling site	State	Lat	Lon	Datum	Z	Code
Thief Lake Wildlife Management Area	MN	48.491	-95.888	WGS 84	14	TI
J. Clark Salyer National Wildlife Refuge	ND	48.612	-100.698	WGS 84	18	JCS
Lake Alice National Wildlife Refuge	ND	48.371	-99.102	WGS 84	13	LA
Beaver Lake Waterfowl Production Area	QN	48.868	-102.629	WGS 84	9	BV
Rush Lake Waterfowl Production Area	QN	48.373	-100.224	WGS 84	10	RU
Sand Lake National Wildlife Refuge	SD	45.793	-98.238	WGS 84	10	SL
North Detroit County	SD	45.783	-98.010	WGS 84	19	NDT
Knutson's Bay	MT	48.388	-104.571	WGS 84	11	KB
Bowdoin National Wildlife Refuge	MT	48.386	-107.665	WGS 84	16	BOW
Benton Lake National Wildlife Refuge	MT	47.674	-111.327	WGS 84	14	BL
Red Rock Lakes National Wildlife Refuge	MT	44.642	-111.841	WGS 84	19	RR
Grays Lake National Wildlife Refuge	Ð	43.059	-111.446	WGS 84	19	GL
Oxford Slough Waterfowl Production Area	Ð	42.253	-112.000	WGS 84	19	SO
Bear River Migratory Bird Refuge	UT	41.486	-112.278	WGS 84	20	BR
Total samples size					208	

Table 2. Sampling locations for each colony included in genetic analyses from the breeding range of the Franklin's gull. The

Table 3. Frequencies for 34 shared mtDNA haplotypes (115 total haplotypes) for 14 populations of Franklin's
gulls sampled throughout breeding colonies found in the United States. Dashes indicated the haplotype was not
found in the population.

	Α	В	ပ	D	н	F	9	H	I	ſ	К	Г	W	N
Thief Lake WMA			•		1					-			-	
Sand Lake NWR	1	ł	a	а		-	5	,	э	•	а	2	3	×
North Detroit County	r	i	Ι	,	7	I	1	,		1	a		-	
Rush Lake WPA	a	•		,	-	r		,		ļ	ı	-		
J. Clark Salyer NWR	1					4	2			•	,	-		a
Lake Alice NWR	e ne	,		x					,	1	a	-	7	
Beaver Lake WPA	1	1	1	ч		r	Ř.	6	c				•	
Knutson's Bay	•		•	а	9	a.	1		а	5	a		1	1
Bowdoin NWR	,	ю	1	1		•	,		1	1	-	-	-	a
Benton Lake NWR	1	a.	•	-	5	1	-		i e	,	r			,
Red Rock Lakes NWR	ı						2	•	э	9	4	-	ę	1
Oxford Slough WPA		,		1	2			1	I		1	6 6	1	S (1
Gray's Lake NWR		a		1	1	1	e		s 1)		,		2	ı
Bear River MBR		a		a	1	•	э	2	ા	1	1	ŝ	1	1
Total Haplotypes Freq.	2	4	4	7	12	Э	5	4	5	2	5	10	13	4
	0	Р	0	R	S	Т	n	Λ	M	X	Y	Z	AA	AB
Thief Lake WMA	-	-					e							-
Sand Lake NWR		1	ä		a		ા	1	а	5	,	ા	•	-
North Detroit County			, i	,			x		,	1		a	1	
Rush Lake WPA			ı	,			r	1				1	•	
J. Clark Salyer NWR		1	,	I	1	1	1	7	1			a.	•	
Lake Alice NWR	,	ı		1	г			,	,	1	-	-	1	a
Beaver Lake WPA	ı	a		•		•	•	r	£	l n		e e		
Knutson's Bay		г	,		а	9	1	,	1		-	1		-
Bowdoin NWR	,	r	,		,		,	9				a	,	3
Benton Lake NWR	ē				e	,		ï	,	ı	¢	x	,	
Red Rock Lakes NWR		ю			а	•	a		1	-	,	,	-	
Oxford Slough WPA	,	e ar			,		a			а	,	а	1	2
Gray's Lake NWR	1	1	2	ę	r		,	,	,	1	•			
Bear River MBR		6	30	1	-	-	0.00	ı.				•	1	١.
Total Haplotypes Freq.	2	10	2	2	3	2	2	4	2	3	2	5	4	4

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	AC	AD	AE	AF	AG	HY	
Thief Lake WMA		4	а	- 1	-		
Sand Lake NWR	a	a	а	a	1		
North Detroit County	2	1	2	ı			
Rush Lake WPA	1	-	ı	ı	т	ľ	
J. Clark Salyer NWR	r		r	r	E	E	
Lake Alice NWR	c	L.	e	c	в	E	
Beaver Lake WPA	т	a		1	4	1.2	
Knutson's Bay	4		,	a			
Bowdoin NWR	a	a	a	a			
Benton Lake NWR	1	а	,	ī	1		
Red Rock Lakes NWR	1	1		-	r		
Oxford Slough WPA		1	ï	r	r		
Gray's Lake NWR	r	Ľ	r	c	r	12	
Bear River MBR	a.	1	Na s	n:	,		
Total Haplotype Freq.	ю	4	2	2	7	2	

Populations	Geneti	Genetic Polymorphism	hism	
	(I)N	(H)N	$h \pm SD$	$\pi \pm SD$
Thief Lake WMA, MN	14	14	1.0000 +/- 0.0270	0.2092±0.1079
J.Clark Salyer NWR, ND	18	16	0.9869 +/- 0.0229	0.28465 ± 0.1436
Beaver Lake WPA, ND	9	9	1.0000 +/- 0.0962	0.0107 ± 0.0074
Rush Lake WPA, ND	10	10	1.0000 +/- 0.0388	0.2556±0.1347
Lake Alice NWR, ND	13	12	0.9872 +/- 0.0354	0.1242 ± 0.0649
Sand Lake NWR, SD	10	6	0.9778 +/- 0.0540	0.1592 ± 0.0853
North Detroit Township, SD (NDT)	19	16	0.9825 +/- 0.0223	0.2763±0.1390
Knutson's Bay, MT	11	11	1.0000 +/- 0.0388	0.4189 ± 0.2201
Bowdoin NWR, MT	16	14	0.9750 +/- 0.0348	0.3124±0.1590
Benton Lake NWR, MT	14	11	0.9341 +/- 0.0611	0.11747 ± 0.0610
Red Rock Lakes NWR, MT	19	14	0.9591 +/- 0.0307	0.1579 ± 0.0799
Grays Lake NWR, ID	19	17	0.9883 +/- 0.0210	0.1607 ± 0.0814
Oxford Slough WPA, ID	19	17	0.9883 +/- 0.0210	0.3182±0.1599
Bear River MBR, UT	20	15	0.9684 +/- 0.0254	0.2082 ± 0.1046
TOTAL	208	182		

Table 4. Franklin's gull sampling locations (mapped in Figure 1), N(I) number of individuals, N(H) number of haplotypes, *h* haplotype diversity, π nucleotide diversity, and SD standard deviation.

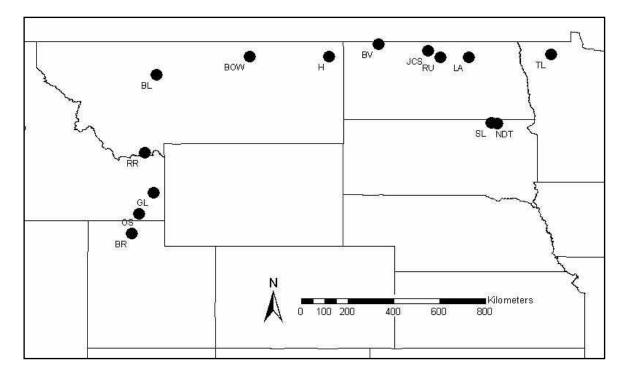


Figure 1. Sampling locations from the breeding range of the Franklin's gull. The letter code corresponds to sampling information (Table 2) for each colony location included in genetic analyses.

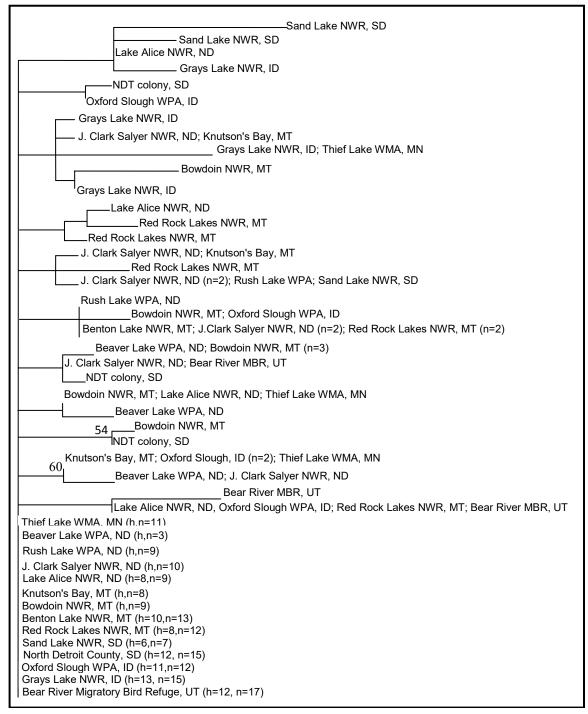


Figure 2. Phylogenetic tree produced using maximum likelihood criteria in GARLI -ln likelihood = 1701.10512. Bootstrap support is given at appropriate nodes. Each branch represents one haplotype. Parentheses prior to nodes indicate posterior probabilities. Parentheses in the resolved portion following location represent number of individuals that share the haplotype represented. The remaining 150 Franklin's gulls compose a polytomy of individuals represented among all colonies sampled. Parentheses following colony location indicate h, number of haplotypes, n, number of individuals.

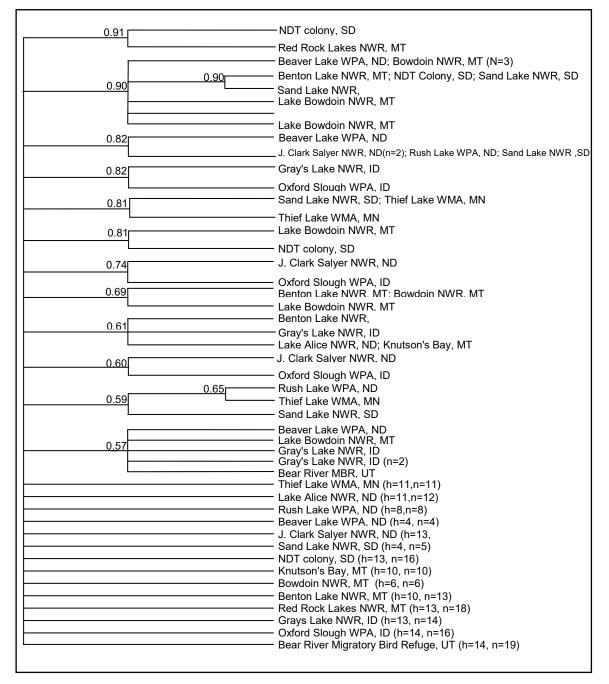


Figure 3. Phylogenetic tree produced using maximum likelihood criteria in program MRBAYES -ln likelihood = 1557.62088. Posterior probabilities assigned to nodes assess the strength of the relationships found at branch nodes. Each branch with resolution represents one haplotype. Parentheses in the resolved portion following location represent number of individuals that share the haplotype represented. The remaining Franklin's gulls sampled compose a polytomy of individuals represented among all colonies sampled. Parentheses following colony location indicate *h*, number of haplotypes *n*, number of individuals.

Source of Variation	df	Variation (%)	FST
Among populations	13	0.60	
Within populations	195	99.40	
TOTAL	208	100	0.00597

Table 5. Results of Analysis of molecular variance (AMOVA) based on Franklin's gull mitochondrial D-loop control region.

CHAPTER III

ASSESSMENT OF HABITAT ASSOCIATIONS OF FRANKLIN'S GULLS IN THE PRAIRIE POTHOLE REGION USING GIS AND REMOTE SENSING TECHNOLOGIES

Abstract

The success of any avian management strategy relies on a firm understanding of factors affecting habitat use. This is of even greater importance when key habitats are limited across the landscape and even more central when substantial numbers of a given species are reliant on only a few sites. Franklin's gulls are one of many colonial-nesting waterbird species designated as a species of conservation concern (e.g. American white pelican, Forster's Tern, Elegant Tern, California Gull; Kushlan et al., 2002), having a large proportion of North America's population concentrated in a select few sites. To assess habitat use of Franklin's gulls within their breeding grounds, I used geographical information systems (GIS) and remote sensing technologies to compare habitat characteristics between wetlands used and unused by Franklin's gulls. Specifically, I created a GIS database and produced digital maps in ESRI ArcGIS 9.3 based on georeferenced Airborne Environmental Research Observation Camera (AEROCam) imagery taken to evaluate land cover at colonial nest sites. Paired t-test results indicate the percent of emergent vegetation land cover was significantly different between sites, being higher in all wetland basins compared to the control wetlands, suggesting the

persistence of Franklin's gulls on wetlands in the PPR depends on a minimum percentage of wetland basins need to be emergent vegetation stands.

Introduction

Wetland degradation and loss impacts many biological organisms (Rahel 2002); most notably amphibians (Cushman, 2006), wetland birds (Johnson, 2001), and aquatic vegetation (Lougheed et al. 2008), which supports many taxa during some point in their life cycle (Mitsch & Gosselink, 1986; Bryan & Scarnecchia, 1992). For example, the prairie pothole region (PPR) possess rich, fertile soils dotted with many wetlands creating both a biologically productive area that provides optimal breeding habitat for wildlife, and simultaneously, a valuable commodity for high-production agriculture (Sugden & Beyersbergen, 1984; Dahl, 1990) which generates a conflict between conservation and economic development.

As a result of economic development, only a small portion of the original PPR wetland base remains due to deliberate drainage (surface and tile) designed to enhance agricultural production (Dahl, 2000). At the time of early settlement (1600s) it is estimated that 90 million hectares of wetlands existed in the U.S. By the mid 1980s less than half (42 million ha) remained in the lower 48 states (Dahl & Johnson, 1991), with more than 90% lost within the Midwestern states (Dahl, 1990). Such losses leave the remaining wetland habitats in great need of protection. Despite federal legislation and policies to protect and restore wetlands (e.g. The Clean Water Act, 1972; Food Security "Swampbuster" Act, 1985; No Net Wetland Loss policy, 1989), wetlands are still being lost at alarming rates (Dahl, 2000). For example between 2004 and 2009 over 25,200 hectares (ha) were lost (Dahl, 2011).

Wetland loss and degradation results in isolated habitat patches, that can have significant impacts on wildlife at the population level (Funk et al., 2010). Management of species possessing small, isolated populations requires identification of essential habitat requirements for continued success. The Franklin's gull is a colonial waterbird species that breeds in large marshes of North America. Because of the ephemeral nature of prairie marshes and their vulnerability to drought and drainage, breeding colonies often shift localities from year to year; in some years local populations are suspected to forego annual breeding opportunities altogether (Burger & Gochfeld, 2009) in response to changing water levels and vegetation structure. Breeding populations have become reduced and segregated over time, likely due to habitat loss and increased disturbance at the breeding sites (Burger & Gochfeld, 1994).

During the Dust Bowl era (1930s) most of the Franklin's gull historic breeding sites were lost due to large-scale drainage projects (Burger and Gochfeld 1994). More recently, the creation of large wetlands, mainly on state and federally protected wildlife refuges, has allowed population expansion. The Franklin's gull breeds almost exclusively on these types of wetlands within the U.S. Even with recent population expansion, there exist fewer than 20 breeding colony locations in the U.S. in a given year, with most of the population concentrated within a portion of the available locations (Burger & Gochfeld, 1994). For example, since the 1900s, the largest Franklin's gull breeding colonies (more than 100,000 breeding pairs) have been located on marshes in the PPR, primarily in North Dakota and Minnesota (Burger & Gochfeld, 1994). Because of their threatened habitats, disjunct populations, and few, but dense nesting colonies, the Franklin's gull is designated a species of high conservation concern in Minnesota, North Dakota, South Dakota, Montana, and Idaho (Table 1).

The Franklin's gull is a long-distance migrant that winters primarily along the western portion of coastal South America with a small isolated inland population in Peru (Burger & Gochfeld, 1994). The combination of decreasing wetland habitats in the contiguous U.S. and few breeding locations makes protection of wetlands critical to the species success. Given that the northern U.S. is at the southern extent of the Franklin's gull breeding range, it is important that habitat sites are maintained. If the trend of wetland loss continues, the Franklin's gull may be forced to breed at higher latitudes, exclusively in Canada, increasing the migration distance and associated survival stressors. Understanding their habitat needs and preferences is essential to maintain quality nesting areas within the U.S.

Previous habitat use studies indicate that the structure and cover of local vegetation may be more important than the plant species composition in selection of desirable sites for breeding wetland-dependant bird species (VanRees-Siewert & Dinsmore, 1996). The Franklin's gull uses aquatic vegetation to build nest structures and as protection from predators (Burger, 1974). The documentation of the amount of emergent vegetation stands, marsh habitat, and other cover types influencing wetland site selection is useful to inform management decision. These factors have not been examined for the Franklin's gull.

I used site-specific geospatial data in the form of aerial imagery taken at the time of nesting to assess the habitat use within Franklin's gull nesting colonies. The objective of this study is to evaluate land cover at established nesting colonies, as well as comparable nearby wetland sites not used by Franklin's gulls using a GIS framework. Two study sites were located in North Dakota, one in South Dakota, and one in Minnesota. These sites were compared to sites with similar wetland characteristics, which were not used as nesting colonies by Franklin's gulls to evaluate habitat selection. Results of this research will provide a better assessment of Franklin's gull occupancy at nesting locations where they are successful. I predicted the percent cover of emergent vegetation and wetland basin size would be the two factors most important to Franklin's gulls for nest site selection. Reasons for my predictions include their site choice is most likely based on needs for nesting, which emergent vegetation is used as building material, stabilization, and cover from predators; a wetland must provide adequate cover and protection which is likely within a preferred range for the Franklin's gull and I predicted larger sites to be occupied with colonies.

Methods

Wetland Habitat Selection

Franklin's gull habitat use was investigated by analyzing wetlands with established nesting colonies ('used' sites) then pairing each wetland basin with a control site with no nesting Franklin's gulls (control sites). I assessed used sites to determine wetland basin size (ha) and class (Table 6, Stewart & Kantrud, 1971). I then used the U.S. Fish and Wildlife Service (USFWS) National Wetlands Inventory (NWI) database to determine control sites. I queried the NWI wetland database to find the basin nearest to each used site, which was of similar size and class (regime). I set the smallest "used" wetland size as the minimum and the largest "used" as a maximum (size in acres). The paired wetland that 1) fell between the smallest and largest wetland used by Franklin's gulls and 2) within the wetland basin size and class restrictions was selected. A GIS with all of these selected wetland basins were sent to the Upper Midwest Aerospace Consortium (UMAC) at the University of North Dakota (UND) in the form of shapefiles for requested image collection. Aerial imagery was taken of four used wetlands (Figure 4) and paired control sites between June and August, 2010.

Study Area

Thief Lake Wildlife Management Area contains 22,240 hectares in Marshall County in northwestern Minnesota and is managed by the Minnesota Department of Natural Resources (MNDNR). Thief Lake itself is Class V (permanent wetland, Stewart & Kantrud, 1971) impoundment of approximately 2,870 hectares, and the surrounding area of the WMA comprises a variety of habitat types (e.g. emergent wetland, conifer and deciduous trees, grasses and hay land, and crop land). The control site for Thief Lake WMA is Nelson Slough found within East Park WMA located approximately 29 kilometers west of Thief Lake WMA. This WMA is approximately 4,220 hectares; it too is an impounded wetland, Class V, managed by the MNDNR. The land cover at this site consists of a mosaic of habitat types similar to the Thief Lake WMA site but with more woody vegetation including conifer and deciduous trees.

J. Clark Salyer National Wildlife Refuge (NWR) is 23,755 hectares and located along the Souris River in north central North Dakota. It is managed by the USFWS primarily for migratory birds species as a breeding or stop-over site for more than 300 species, which contributed to its designation as a Globally Important Bird Area (GIBA) by the U.S. IBA Technical Committee. The Refuge encompasses a high diversity of habitat types from prairie and riverine to forest systems. Common management of these diverse ecosystems includes prescribed fire and intensive water-level management. In 2010, the wetland impoundment (Class V) at approximately 610 hectares, with nesting Franklin's gulls was located in McHenry and Bottineau counties between Dam # 326 to the north and Dam #320 to the south. This nesting site was compared with Round Lake, a Class V lake at approximately 1,010 hectares, which was located approximately 40 kilometers southeast of the J. Clark Salyer wetland in Pierce County.

Lake Alice NWR is a 4,650-hectare habitat complex, comprised mainly of wetland, marsh, and grassland habitats. It is located in Ramsey and Towner counties, North Dakota. Continuous flooding from the nearby Devils Lake basin challenges optimal water level management for wetland-dependant wildlife at this Refuge. Despite periodic flooding, Lake Alice NWR regularly hosts one of the world's largest Franklin's gull breeding colonies in the U.S. (Burger & Gochfeld, 2009). During the nesting season of 2010, the colony breeding on this Class V wetland utilized much of the northern portion of the Refuge. Cranberry Lake was used as a comparison site for Lake Alice NWR. It too is a Class V lake, 930 hectares in size, and is located approximately 48 kilometers southwest of the Lake Alice pool in Benson County.

Sand Lake NWR is 8,620 hectares of diverse wildlife habitat, located in Brown County in northeastern South Dakota. The James River runs through this area allowing refuge management to manipulate water levels, optimizing wetland objectives for fish, wildlife, and recreation. The Refuge impoundment that is used by nesting Franklin's gulls is, Mud Lake, a Class IV-V wetland, approximately 2,145 hectares in size. Sand Lake NWR is listed as a GIBA and is also designated as a Wetland of International Importance by the Ramsar Convention (Convention on Wetlands of International Importance 1971). In 1994, Sand Lake NWR hosted the world's largest breeding colony of Franklin's gulls with 150,000 breeding nesting pairs (USFWS 2005). Sand Lake NWR was paired with Lake St. John, a 565 hectare, Class V wetland located approximately 145 kilometers southeast in Hamline County.

Imagery Acquisition and Processing

UND developed an airborne multispectral digital imaging system in 2001 called Airborne Environmental Research Observational Camera (AEROCam). AEROCam was developed through a unique partnership with several UND departments, including the UMAC, the School of Engineering & Mines, and the flight operations at the John D. Odegard School of Aerospace Sciences. UMAC is led by UND, and covers the states of North Dakota, South Dakota, Montana, Wyoming, and Idaho with partners from academia, industry, and the government to provide services for farmers, ranchers, educators, researchers, and natural resource managers. Capabilities of AEROCam include providing imagery taken within a desired time frame in visible and near-infrared bands at a higher resolution than can be offered by alternative imagery sources in a short period of time and at no cost to selected users.

The imaging system includes a Redlake MS4100 area-scan multi-spectral digital camera that features a 1920 x 1080 CCD array (7.4-micron pixels), with 8-bit quantization. Images were delivered in TIFF format along with tabular flight, camera, and GIS data with the approximate GPS center of each image. Image processing was performed in Leica Photogrammetry Suite and ERDAS IMAGINE 2010 (Norcross, GA). All images were registered to Universal Transverse Mercator (UTM) projections, North American Datum1983 (NAD 83), with corresponding zones 12-15 (north). Digital

Elevation Models (DEM) were used at a 30-m resolution provided by NASA's Shuttle Radar Topography Mission (SRTM), downloaded from the U.S. Department of Agriculture and Natural Resource Conservation Service (NRCS) website Datagateway (http://datagateway.nrcs .usda.gov/). DEMs provided vertical information required for triangulating ground control points (GCPs) during referencing. The most recent National Agriculture Imagery Program (NAIP) orthophotos available (2009 or 2010) were used to provide the horizontal referencing information needed in triangulation. NAIP images were also acquired from the USGS Datagateway.

Sites were flown in lines from north to south and south to north to capture the full extent of each study area including some overlap for ease in the georeferencing process. It typically took three to five flight lines per wetland site to capture enough aerial images to cover the site. Each of these flight lines had a range of six to 40 images. During image processing, clearly identifiable GCPs (e.g. road and dike intersections, trees, permanent structures) with associated geographic coordinates were selected in each image. A minimum of 6-8 control points per corner per image were used. Once each flight line had all GCPs, triangulation was performed before moving on to the subsequent flight line images.

To determine referencing accuracy, a root mean squares error (RMSE) report was computed to measure the differences between our predicted value of 1-m resolution and the values of the control points within the flight line. Maximum RMSE of rectification was calculated for each site. Calculated RMSE values were used to characterize the strength of a registration and measure spatial accuracy. Lower RMSE values suggest greater spatial accuracy; the predicted data are closer to the observed data (Grapentine, 2010). A 2.0 RMSE was chosen given the difficulty of referencing wetlands and the challenges of finding quality and quantity GCPs and was acceptable for obtaining the level of detail needed for assessing land cover. All flight line images had an accuracy measure of < 2.0 RMSE from the triangulated GCP's before being accepted as accurate and processing as an orthorectified image. Orthophotos were then mosaiced into a single TIFF image file. Georeferenced AEROcam images were uploaded and archived on the Digital Northern Great Plains (DNGP) geospatial data archive (www.dngp.umac.org).

Photo Interpretation

I followed land use classification categories as guidelines to determine classifications, descriptions of wetland basins, and land directly adjacent to the basin using the USGS National Land Cover Database (Homer et al., 2004, Table 7). This classification was a newer version using the U.S. Department of Agriculture (Anderson et al., 1976) classification system. The data were collected at a low altitude (< 3,100 m), which allowed determination of land-cover categories. Shapefiles were created in ArcGIS 9.3 for each land cover category (layer) and digitized at appropriate scales to delineate boundaries of habitats. Habitat boundaries were digitized based on specific color and texture patterns in the orthorectified mosaic. Cover class color schemes and coding followed the USGS Level I land use color code (Anderson et al. 1976) to represent land cover corresponding to habitat types. Examples include: water is dark blue, wetland is light blue, barren land is gray, agricultural land is brown, and rangeland is light orange. Colony sites were delineated by the natural wetland basin perimeter or by the impoundment of water in which the colony was located. For example, if a wetland was managed separately by intentionally controlling the water levels in a designated area, the

boundary of the colony site would be delineated by the structures (dams, road, etc.). In natural, un-impounded wetland basins, the natural perimeter of the wetland basin was determined as the colony site. Land cover classification was restricted to the area directly surrounding the colony sites and control sites (30 m buffer) to determine land use, specifically presence of crop and roads as this may introduce disturbance caused by humans, vehicles, or machinery noise. Also, during the breeding season Franklin's gull activities remain within 30m of the colony (Beyersbergen, 2004).

Land Cover Analysis

To analyze the resulting database, GIS techniques assessed the heterogeneity of land-classification parameters (Jain et al., 2010). ArcGIS 9.3 (ESRI, 2009) was used to create and analyze land-cover layers. Referenced images of wetland basins were classified into various land-cover categories and used to quantify and compare the following variables: 1) cover type within the colony and in surrounding wetland habitats, 2) percent emergent vegetation, 3) basin size (ha), 4) edge to water ratio, and 5) presence of roads around the wetland perimeter. The cover type within the wetland sites was investigated to determine if there was a preference between vegetative characteristics at sites with colonies compared to wetlands with similar features. Emergent vegetation is known to be an important component to the nesting phase in Franklin's gull breeding sites as they are dependent on vegetation for nest building. Investigating the difference between sites would provide valuable information on the quantity preferred at nesting locations. Determining if there is a size criterion among wetland basins is important information since most of the nesting colonies are on managed wetland impoundments. If wetland basin size is an important variable used in site selection for nesting Franklin's

gulls it is possible to manage water levels during the breeding season to provide desired habitat at that specific time. Many species are sensitive to edge effects, as more edge increases the chance of predation, disturbance, and lesser quality habitat (Yahner, 1988). Investigating the relationship among edge to water ratios of wetland basins will provide information on sensitivity levels of edge to colonies. Disturbance is another variable analyzed at these sites in the form of roads to determine if the presence or the quantity is affecting colony site locations. Examination of these five variables will provide an understanding of what habitat requirements are needed by Franklin's gull to select a site for breeding colonies.

After digitizing was complete, each of the land-cover categories (layers) was calculated (Table 8) and converted from area (ha) to percent cover of the wetland basin (Table 9) for comparison. Characteristics of the wetland sites were examined using the mean and standard deviations along with paired t-tests to determine if there was a difference between land-cover types for breeding sites versus control areas. For each land cover category the four control sites were averaged as were the four used sites to obtain mean and standard deviation calculations to objectively determine the difference among land-cover categories between sites with and without Franklin's gull colonies. A paired t-test was performed among the landscape characteristics to determine significant differences ($\alpha = 0.5$) between wetlands with nesting Franklin's gull colonies and without nesting colonies. Small sample size can influence the power of the *t*-test therefore alpha values up to 0.15 were considered and may indicate a trend for the data described in this study (if $p \le 0.15$).

57

Results

Detecting a significant difference from mean and standard deviation results was hard to interpret, however they indicate there may be a difference among some landcover variables at wetland sites with and without Franklin's gull nesting colonies (Table 10). Paired *t*-test results revealed a significant difference between the amount of emergent vegetation at wetland nesting sites (p=0.03) compared to sites that did not have nesting colonies of Franklin's gulls (Table 11). Relaxing the alpha value for small sample size to 0.15 detects a significant difference for the amount of wetland edge ("perimeter", p=0.14) and the amount of marsh habitat between wetland sites (p=0.12, Table 11). Among wetlands with Franklin's gull colonies (J. Clark Salyer NWR, Sand Lake NWR, Thief Lake WMA, and Lake Alice NWR), all had emergent vegetation dispersed throughout the wetland basin, not exclusively at the perimeter as seen in their paired control wetland basins (Appendix B). Breeding sites also had a dominance of certain land-cover types within all the wetland basins; all had a combination of open water, deep marsh emergent vegetation stands (cattail/bulrush), and/or marsh habitat that comprised almost the total percent land cover (> 94%). Although present, non-aquatic vegetation species (woody, herbaceous, and crop) as well as exposed ground contributed little to the percent cover (<3% average across sites).

Discussion

I predicted that the amount of emergent vegetation and wetland basin size would be key variables which influenced the occupancy of wetland basins by nesting Franklin's gulls. The results from this study found the amount of emergent vegetation was significantly different compared to the paired (control) wetland basins; however wetland

basin size along with all other variables measured did not show a significant difference between sites. Because of the complexity of these wetland systems and specific nesting needs of Franklin's gulls, it seems that multiple variables may be contributing to wetland site selection for nesting (implications from this study, Burger and Gochfeld, 1994, Burger 1974). An alternative approach may be to quantify not only wetland cover classes, but also surrounding cover across the landscape to determine presence of alternative nesting sites. Examination of wetland complexes, rather than individual wetland sites may be more appropriate for a species like the Franklin's gull; in this species, wetland use for rearing young is not based on site fidelity but rather an opportunistic event based on characteristics of ephemeral habitats. Habitat heterogeneity for wetland dependant bird species has been recognized as the basic component to increasing waterbird species diversity (Fairbairn & Dinsmore, 2001). Though not significant, data indicated that a higher portion of nesting colonies occurred on sites with large amounts of open water and emergent vegetation, or at sites where open water and marsh habitat dominated the cover type. These combined factors may indicate that the presence of hemi-marsh (Weller & Spatcher, 1965) may be the key factor in Franklin's Gull habitat use. Other possible colony site selection factors (e.g., degree of wetland isolation) are likely tied to a larger geographic scale. Understanding local factors, as was the case in this investigation, is important for managing and conserving individual wetlands, but larger-scale perspectives are critical for understanding and managing populations in fragmented landscapes.

This project had several inherent limitations that had a notable effect on the significance (or lack thereof) of these results. Franklin's gull nesting sites in the U.S. rarely exceed 20 locations (Burger & Gochfeld, 1994); therefore, sample size was limited

for the habitat analysis portion of this study. Sample size was further limited based on limited resources (e.g. imagery taken by AeroCam was not received georeferenced, as a result additional sites were unable to be manually referenced for difficulty of obtaining adequate GCPs in water images; funds and time to support manually referencing images were insufficient). Lack of identifiable ground features and abundant open water made georeferencing of some sites difficult (Grapentine & Kowalski, 2010). In addition, spring flooding occurred throughout the PPR during 2010 further reduced the number of nesting colonies (n = 7 pairs of breeding colonies and associated controls). Another factor that decreased our sample size was changes in colony nest locations of basins which were predetermined for aerial imagery acquisition, which reduced the final sample size to four sites and their paired control wetlands.

The information provided by the AEROCam aerial images captured at colonial nest sites for the Franklin's gull provided a valuable assessment of habitat features present during colonial nest site selection for the Franklin's gull compared to other sources of aerial imagery which are not captured at the time of nesting for this species. Additionally, the resolution of the imagery acquired was 1-m, which allowed a much more detailed and accurate assessment of the land cover, quantifying the variables in this study with more precision and accuracy compared to other sources of remotely sensed aerial imagery at no cost. Despite the limitations imposed by small sample sizes, trends in results provided valuable information for wetland managers.

In conclusion, I recommend that the above approaches to investigating habitat use of Franklin's gulls be further developed. Future work should be adapted to include measuring interspersion (water-vegetation) metrics, and should also consider issues

encountered during wetland geo-referencing such as shifts in nesting locations and ability to reference landscapes without identifiable ground control points (water). To ensure that quality habitat is provided for breeding, a myriad of habitat features are necessary: emergent vegetation representing a hemi-marsh condition (Weller & Spatcher, 1965), appropriate water levels, and more importantly, is the proper timing for the occurrence of these ephemeral features. Franklin's gull colonies have been known to have multiple species intermixed during nesting such as White-faced Ibis (Plegadis chihi), Pied-billed Grebe (Podilymbus podiceps), American Coot (Fulica americana), and Redhead (Aythya *americana*) (Burger, 1974). Thus, the management for breeding colonies of Franklin's gulls will provide suitable habitat for these other over-water nesting, wetland-dependent birds. Further, the selection of wetland sites changes between years based on fluctuating water levels and vegetative cover (Burger, 1974). Habitat features should be managed to meet these needs at a landscape level (multiple basins, wetland complexes) at the time of breeding pair arrival on the site. A comprehensive ecosystem approach will protect vital habitat for Franklin's gulls and other wetland species, maintaining the heterogeneity of these biological systems.

Remote sensing and the resulting images are important tools to provide information for decisions in wetland management by elucidating the role of individual biotic and abiotic factors. This study aimed to investigate habitat parameters found at nesting locations during the nest initiation stage through quantifying real-time land-cover data. Though the scope of this study is limited by the small sample size of nesting (used) and corresponding number of control sites, the results provide trends that can be used to better manage wetlands for use as Franklin's gull nesting sites.

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Table 6. Description of wetland classifications measured to evaluate Franklin's gull use of wetland basins in the PPR of the U.S. in 2010; adopted from Stewart & Kantrud, 1971.

Wetland Class	Wetland Name	Description
Ι	ephemeral ponds	The wetland-low-prairie zone dominates the deepest part of the pond basin.
II	temporary ponds	The wet-meadow zone dominates the deepest part of the wetland area. A peripheral low-prairie zone is usually present.
III	seasonal ponds and lakes	The shallow-marsh zone dominates the deepest part of the wetland area. Peripheral wet-meadow and low-prairie zones are usually present.
IV	semi permanent ponds and lakes	The deep-marsh zone dominates the deepest part of the wetland area. Shallow-marsh, wet-meadow, and low- prairie zones are usually present, and isolated marginal pockets of fen zones occasionally occur.
V	permanent ponds and lakes	The permanent-open-water zone dominates the deepest part of the wetland area. Peripheral deep-marsh, shallow- marsh, wet-meadow, and low-prairie zones are often present, and isolated marginal pockets of fen zone occasionally occur.
VI	alkali ponds and lakes	The intermittent-alkali zone dominates the deepest part of the wetland area. Peripheral shallow-marsh, wet-meadow, and low-prairie zones are usually present. A deep-marsh zone is normally absent except occasionally for isolated patches near marginal seepage areas. A few isolated pockets of fen zone are normally present along the margins.
VII	fen (alkaline bog) ponds	The fen zone dominates the deepest part of the wetland area. Peripheral wet-meadow and low-prairie zones are often present.

Table 7. Description of land cover classifications measured to evaluate Franklin's gull
use of wetland basins in the PPR of the U.S. in 2010 (Homer et al., 2004).

Land Cover Class	Description
Open water	Wetland area dominated by open water > 75%
Emergent herbaceous wetland	Perennial herbaceous vegetation is >80% of vegetative cover and soil or substrate is periodically saturated or covered with water; usually found in dense stands
Marsh - aquatic bed	Intermittent area of wetland between open water and emergent vegetation stands; dominated by plants that grow and form a continuous cover on at the surface water, water <25% cover
Upland - herbaceous vegetation	Areas dominated by graminoid or herbaceous vegetation
Woody	Areas dominated by woody vegetation
Crop	Areas used for production of annual crop; all land being actively tilled
Bare land & soil	Areas with little to no vegetation, exposed land is typically bedrock, soil, accumulations of earthen soils
Roads	Roads including paved, unpaved, and two-track trails
Boundary	Perimeter of wetland basin

		Perimeter	Open	Emergent			Bare			Roads
Site	Basin	(m)	water	Stands	Marsh	Woody	ground	Upland Crop	Crop	(m)
Thief Lake	2820.0	22147.0	0.0	251.0	2569.0	0.0	0.0	0.0	0.0	0.0
Thief Lake Control	877.7	15817.0	179.0	42.3	649.0	7.3	0.0	0.0	0.0	4034.0
Mud Lake	1603.0	19657.0	921.4	266.8	415.2	0.0	0.0	0.0	0.0	3549.7
Mud Lake Control	568.0	10293.0	534.5	1.7	0.0	17.7	0.0	4.2	8.5	2963.0
Lake Alice	5010.0	46342.0	4510.0	252.2	30.7	14.0	37.8	26.2	140.4	0.0
Lake Alice Control	920.5	15635.0	869.5	0.0	11.4	11.8	0.0	27.6	0.0	0.0
J.Clark Salyer	1839.0	17423.0	359.6	44.1	1350.0	6.9	0.0	67.4	6.3	2985.0
J.Clark Salyer Control 1033.0	1033.0	14063.0	1023.7	2.2	0.8	1.5	0.0	1.2	3.5	291.0

Table 8. Land cover (ha) by site of four wetland basins with Franklin's Gull nesting colonies paired with four non-use

	OPEN				BARE		
Site	WATER%	EMERG%	MARSH%	EMERG% MARSH% WOODY%	GROUND %	GROUND % UPLAND % CROP %	CROP %
Thief Lake	0	6	91	0	0	0	0
Thief Lake Control	20	5	74	1	0	0	0
Mud Lake	57	17	26	0	0	0	0
Mud Lake Control	94	0	0	m	0	1	1
Lake Alice	90	5	1	0	1	1	ю
Lake Alice Control	94	0	1	1	0	3	0
J.Clark Salyer	20	7	73	0	0	4	0
J.Clark Salyer Control	66	0	0	0	0	0	0

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Table 10. Characteristics of wetland study sites: mean percentages (\pm standard deviation), mean road length, mean cover to water and edge to water ratios from wetland basins with (used) and without (unused) Franklin's gull nesting colonies across the PPR of the U.S. in 2010.

	Used	Unused
Ν	4	4
Size (ha)	2818.0 ± 1553.4	850.0 ± 199.0
% Open water	42.0 ± 40.0	77.0 ± 37.8
% Emergent vegetation		
stands	8.2 ± 6.2	1.3 ± 2.3
% Marsh/aquatic bed	47.8 ± 41.8	18.8 ± 36.7
% Woody vegetation	0.0 ± 0.0	1.3 ± 1.3
% Bare ground/soil	0.2 ± 0.4	0.0 ± 0.0
% Upland	1.0 ± 1.7	1.8 ± 1.3
% Crop	0.8 ± 1.3	0.5 ± 0.7
Roads (m)	1634.0 ± 1900.0	1822.0 ± 1988.0
Mean cover: water ratio	1295 : 1448	177:652
Mean edge: water ratio	26392 : 1448	13952 : 652

Table 11. Paired t-tests results for land cover characteristics comparing wetland sites with and without nesting Franklin's gull colonies during the 2010 breeding season across the PPR.

Variable	df	t	p-value
Perimeter (edge of wetland basin)	3	2.00	0.14
Open water		0.82	0.47
Marsh		2.14	0.12
Emergent vegetation		3.73	0.03
Woody vegetation		-0.84	0.46
Bare ground		1.00	0.39
Upland/herbaceous vegetation		0.89	0.44
Crop		0.94	0.41
Roads		-0.13	0.90

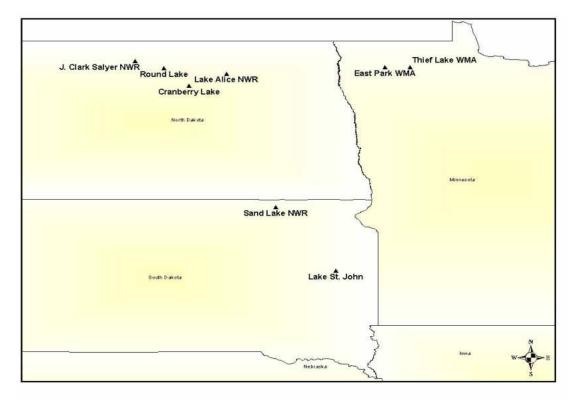


Figure 4. Wetland location sites with established Franklin's gull colonies (n=4) paired with control sites (wetlands, n=4).

CHAPTER IV

CONCLUSIONS

The Franklin's gull is one of many colonial-nesting waterbird species identified as a high priority species for conservation throughout their range. Nesting colonies are located at only a handful of locations throughout the U.S.; these are geographically separated from one another. Conceptions prior to this study were isolated nesting locations may prevent gene flow and limit genetic diversity (Avise, 2004). Isolation of these populations is also a symptom of a second key issue: ongoing habitat fragmentation. It is necessary to provide adequate habitat to maintain genetic diversity for any species of concern. Identifying and quantifying desirable habitat traits for a species is critical to sound management. A combination of both genetic and GIS approaches provided unique insights into the ecology of Franklin's gulls.

This research was the first attempt to assess relationships of Franklin's gull colonies breeding across the U.S. through DNA analysis and additionally quantified habitat use at sample colonies. Results of this investigation indicate that Franklin's gull nesting colonies in the Midwestern portion of the U.S., although geographically segregated, are panmictic. In addition my research shows high levels of genetic diversity among individuals in all colonies sampled. Given the levels of genetic diversity documented in this study there is less concern in the event of colony abandonment or collapse; if suitable habitat is made available in subsequent years shortly following the event of non-use they will return and genetic diversity will remain stable.

Because composition of individuals within colonies varies from year to year, management should focus on ecosystem strategies. In the U.S., Franklin's gull nesting colonies occur almost exclusively on state and federal wildlife areas. This is beneficial to the Franklin's gull given most wetlands are impoundments with abilities to manage habitat at optimal locations and times therefore increases the chance of successfully establishing a nesting site. In any given year there are typically less than 20 colonies in the United States. High levels of mating interaction among individuals, which was revealed through genetic analysis, indicates a low level of natal site fidelity, which is likely due to the ephemeral nature of wetlands in the PPR. Thus, it is important to ensure adequate habitat, though composition of breeding pairs will change from year to year.

The driving force of colony site selection is unknown, but is clearly important. I did not find significant evidence to predict nesting locations from each year, there were identifiable trends. Results of this study suggest that a balance of open water, marsh habitat, and emergent vegetation stands is desirable for Franklin's gull breeding pairs. We know hemi-marsh (equal amounts of emergent vegetation and water in an interspersed pattern) wetlands support high numbers of bird species diversity (Weller & Spatcher, 1965; Rehm & Baldassarre, 2007), likely to enhance prey diversity such as invertebrates for Franklin's gull and other species of wetland birds. It is important for future studies to further quantify these habitat parameters for the success and proper management of Franklin's gulls given that this study failed to detect significance for individual characteristics involved with wetland site selection.

72

This study, while limited in scope, provides an important starting point for management of Franklin's gulls, and serves as a model for integrating information from population genetic and geographic approaches. Additional research is needed to determine whether genetic structure of the remaining breeding range (i.e. Canada) corroborates this study's findings. Further, genetic studies of Franklin's gulls on the wintering grounds would provide valuable insights into the breeding biology of these colonial waterbirds. Information about pair formation as well as determining recruitment factors for spring migration would allow a total evidence approach to ensuring the presence of the species in U.S. ecosystems.

Literature Cited

- Avise, J.C. 2004. Molecular Markers, Natural History, and Evolution, 2nd Edition. Sinauer, Sunderland, MA. 684 pp
- Rehm, E. M., and G. A. Baldassarre. 2007. The influence of interspersion on marsh bird abundance in New York. Wilson Journal of Ornithology 119:648–654.
- Weller, M. W., and C. S. Spatcher. 1965. Role of habitat in the distribution and abundance of marsh birds. Iowa State University of Science and Technology, Ames, Iowa, USA.

APPENDICES

Appendix A Permits

Federal Permit: United States

DEPARTMENT OF THE INTERIOR			3-201
FISH A WILDLEPE U.S. FEH AND WILDLEE SERVICE			(1/97)
		2. AUTHORITY-STATUTES 18 USC 703-712	
FEDERAL FISH AND WILDLIFE PERM	ur	10 000 102112	
U.S. Fish and Wildlife Ser Migratory Bird Permit Off			
P.O. Box 25486, DFC (6015	4)		
Denver, Colorado 80225-0	0486		
(303) 236-8171		REGULATIONS 50 OFR Part 13	
1. PERMITTEE		50 CFR 21.23	
KATHERINE ROSE MEHL		A A A A A A A A A A A A A A A A A A A	
UNIVERSITY OF NORTH DAKOTA			
ATTN: KATHERINE MEHL			
10 CORNELL STREET		3. NUMBER	
GRAND FORKS, ND 58201-9019		MB16962A-0	
		VI YES	5. MAY COPY
		A NO	
		6. EFFECTIVE	7. EXPIRES
		07/28/2010	03/31/2013
6. NAME AND TITLE OF PRINCIPAL OFFICER (X 81 is a buttiness)	9. TYPE OF PERMIT	and the second se	
	SCIENTIFIC COLLECTING		
10. LOCATION WHERE AUTHORIZED ACTIVITY MAY BE CONDUCYED			
MINNESOTA, NORTH DAKOTA, SOUTH DAKOTA, MONTANA, IDAH	O. NEVADA UTHA AND OREGON		
		•	
11. CONDITIONS AND ALITHORIZATIONS			
A GENERAL CONDITIONS BIT OLD HISBART DOFS OFFI 13. AND SPECIFIC CONDITIONS CONTINUED UN MORE APART OF THIS FERMINT, ALL ACTIVITIES AUTHORIZED THEREM HIST BE GARRED CLI FALACORI BASMITTED, CONTINUED VALDITY, OR RENEWAL, OF THIS FERMIT IS BASLECT TO COMPLETE AND THEIR FILMID OF ALL REQUIRED INFORMATION AND REPORTS.	EDERAL REGULATIONS CITED IN BLOCK #2 ABOVE, A D WITH AND FOR THE PURPOSES DESCRIPTION THE	RE HEREBY	
SUBMITTED, CONTINUED VALIDITY, OR RENEWAL, OF THIS PERMIT IS BULLECT TO COMPLETE AND TIMES FILING OF ALL REQUIRED INFORMATION AND REPORTS.	LY COMPLIANCE WITH ALL APPLICABLE CONDITIONS, N	SUUDING THE	
B. THE VALIDITY OF THIS PERMIT IS ALSO CONDITIONED UPON STRICT OBSERVANCE OF ALL APPLIC	CABLE FOREIGN, STATE, LOCAL OR OTHER FEDERAL	1 4141	
C. VALID FOR USE BY PERMITTEE NAMED ABOVE.	-	- ense.	
and the following subpermittees are authorized: Annmarie Krmpotich and	Linha Caulti		
and the resoning suspensitions are automized. Annihilane Kimpouch and	John Cavitt.		
D. You and subpermittee(s) are authorized to take and possess the follow	wing for scientific purposes:		
twenty-four hundred total (2400) Franklin's Gulls (Larus pipixcan) feathers			
*Note: Colony disturbance in <u>Nevada</u> must be kept to a minimum, permit research disturbance	ttee and subpermittee(s) must report i	anv nest abandooment l	hat le coursed hu
research disturbance.	,	any noor abandonmont t	hat is caused by
E. All of the above species and numbers are the totals for the term of request with justification to amend the parmit must be submitted to the less	the complet if a shares to set to		
request with justification to amend the permit must be submitted to the last	und office for additional authorization	the above authorized ac	tivities, then <u>a written</u>
F. You are authorized to salvage migratory birds. Any dead bald eagle or Repository at (303) 287-2110 and to the migratory bird permit issuing offi specimens.	golden eagle salvaged must be reporting at (303) 236,9171 The Reserve	ted within 48 hours to th	e National Eagle
specimens.	at (505) 230-6171. The repositor	y will provide directions f	or shipment of these
G. You are authorized to solve a shandoned (unaccurring) misselves bio			man a con un
G. You are authorized to salvage abandoned (unoccupied) migratory bird eagles or golden eagles, or species listed as threatened or endangered un	I nests and nonviable eggs after the ne	esting season, except for	r nests and eggs of bald
For a list of threatened and endangered species in your state, visit the U.S. www.fws.gov/endangered.	Fish and Wildlife Service's Threater	ed and Endangered Spe	cies System (TESS) at:
THAT TO GET STRUCTURE			
ADDITIONAL CONDITIONS AND AUTHORIZATIONS ALSO APPLY			
2. REPORTING REQUIREMENTS			
ANNUAL REPORT DUE: 01/31			1
1			
SQUED BY			DATE
Kelly Karl & CHEF, MBPO, REGION	6		07/28/2010
augur vera my			STEDIEUTU

State Permit: South Dakota

//	State of South Dakota	
11	Department of Game, Fish, an	nd Parks
Whom It May Concern		License Number_48
his Permit Authorizes:		
Katherine Mehl		
10 Cornell Street	18	
10 Cornell Street		
Grand Forks	ND	58202-1990
		ZIP
llecting for (Institution or Associ		
University of North Dako	ta	
10 Cornell Street		
ADDRESS		
Grand Forks	ND STATE	58202-1990
		ZIP
take, possess, transport, collect,	or study for scientific purposes	the following wild animals in such man
d under such conditions set forth b		
COMMON NAME & SPECIES	NUMBER	VICINITY OF COLLECTION
Franklin's Gull (Larus pipixcan)	1-3 feathers from 20-50 nests	Sand Lake National Wildlife Refuge
		2.

Frior to contecting any bat species, permittees MUST contact the Natural Heritage Program (605-773-4229).

DISPOSITION OF SPECIMENS AND/OR SPECIAL CONDITIONS: Cast feathers will be collected passively

from nests within the breeding colony. The permittee will use a canoe to access the nesting colony. Care will be taken to reduce disturbance to the nesting colony. Blood will be taken from the quill of collected feathers for DNA analysis. Isotope analysis will be

conducted on the vanes of collected feathers.

NOTICE: A copy of this permit must be carried when exercising its authority. Collecting that may be authorized under this permit does not relieve the permit holder from compliance with any Federal law or regulation. ***** The taking of any federal or state threatened or endangered species is prohibited, unless specifically authorized by a state or federal permit. Please inform this office (605-773-4345) if you incidentally take any of these species. ***** The permit holder MUST notify the local Conservation Officer prior to engaging in any collections. The enclosed collection report forms must be submitted to the Department of Game, Fish, and Parks, 523 E. Capitol-Foss Bldg., Pierre, SD 57501 no later than January 31, 2011. This permit is granted under the provisions of SDCL 41-6-32. <u>Permits will expire on the 31st day of December for the year issued, unless a specific collection period is specified.</u>

12

Dated at Pierre, South Dakota this 25th day of June, 2010.

Allo Kompone South Dakota Department of Game, Fish, and Parks By_

License #: GNF02750301 issued 01/29/20 Birthdate: 09/09/1967	010	Resident License(s)
Phone: 701-777-3699 Sex: Female Height: 5 ft 9 In Weight: 130 Hair: Brown Eyes: Hazel		Scientific Collection (Expires 12/31/2010) Species: All species except Bald Eagles/Endangered/Threatened Species (mist netting and bandings for ornithology class) Mallard (200), Blue-Winged t (100) and Bobolink (200)
KATHERINE MEHL 10 CORNELL ST GRAND FORKS ND 58	202	2011211.61
Report all poachers 1-800-472-2121	http://gf.nd.gov	Licensee Signature Nontransferable/Nonrefunc

License #: GNF02750301 Issued:01/29/2010

State Permit: Minnesota

STATE OF MINNESOTA DEPARTMENT OF NATURAL RESOURCES DIVISION OF ECOLOGICAL RESOURCES 500 LAFAYETTE ROAD, BOX 25 ST. PAUL, MINNESOTA 55155-4025

SPECIAL PERMIT NO. <u>16509</u> (scientific collection)

March 20, 2009

To Whom It May Concern: Permission is granted to:

Katherine Mehl Assistant Professor Department of Biology University of North Dakota 10 Cornell St. Grand Forks, ND 58202-9019 701-777-3699

For the purposes of research, to passively collect and to possess Franklin's Gull (*Larus pipixcan*) feathers in Minnesota subject to the following conditions:

- 1. The Minnesota trespass laws apply for all activities on private land;
- Permittee shall be solely responsible for any and all damage or injury to persons, domestic or wild animals and real or personal property of any kind, resulting from the activities undertaken pursuant to this permit;
- Permittee shall hold the Department of Natural Resources, its officers, agents, and employees harmless from any and all liability and damages resulting from any activities undertaken pursuant to this permit;
- Permittee may authorize subpermittees, provided they retain a copy of this permit in their possession while conducting permitted activities;
- Reasonable precautions are to be taken to keep disturbance of birds to a minimum during collection and sampling of passerines;
- A report of activities carried out under this permit including the number of birds banded and sampled is to be submitted to the DNR's Division of Ecological Resources, Box 25, 500 Lafayette Road, St. Paul, MN 55155 Attn: permits, by January 15, 2011.
- 7. This permit is effective immediately through December 31, 2010, but may be revoked at any time.

Lori N. Naumann Division of Ecological Resources Special Permits

Cc: Captain Ken Soring, Regional Enforcement Supervisor Jeff Lightfoot, Regional Wildlife Manager Lindsey Peterson, Wildlife Research Maya Hamady, Regional Nongame Wildlife Specialist Katie Haws, Regional Nongame Wildlife Specialist Captain James Dunn, Regional Enforcement Supervisor Paul Telander, Regional Wildlife Manager Elizabeth Roberts, US Fish and Wildlife Service

State Permit: Utah

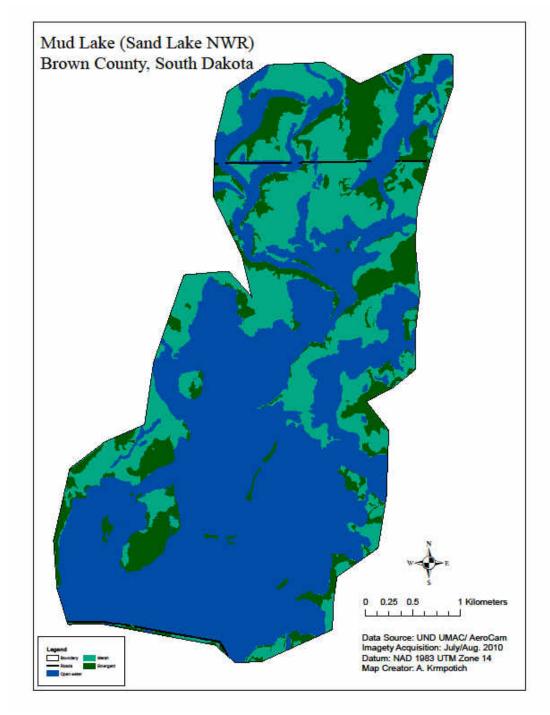
	CERTIFICATE O	FREGISTRATION
Registrant Name and Address KATHERINE MEHL UNIVERSITY OF NORTH DAKOTA 10 CORNELL ST.		COR Number 2COLL8521
GRAND FORKS ND 58202-9019		1
Phone (701) 777-3699		Fee Received \$85.00
		Effective Dates: From: 8/26/2010 To: 7/31/2011
COR Type: COLLECT/POSSESS	(USFWS PERMIT NO. MB16962	21-0) Activities Report Due Data: 8/31/2011
BE OBSERVED. Specific Provisions: AUTHORIZ	ED TO COLLECT FEATHERS F	ON, HEALTH AND CARE OF ANIMALS USED IN VARIOUS STUDIES MUS ROM THOSE SPECIES OF BIRDS IDENTIFIED HEREIN AND FROM IZED WITHIN THE BOUNDARIES OF DIVISION RIPARIAN BIRD SURVEY
SUBPERMITTEES: JOHN CAVITT, AN		
COLLECTING REPORT OF ACTIVITIE STUDY RESULTS ARE DUE UPON TH		CONTACT (FORMS ENCLOSED) DUE BY AUGUST 31, 2011. Y.
COR shall be in possession of registrant	Istrant must be recorded immedia	tely to the Wildlife Begistration Office 1594 W N Temple Suite 2110 Box
Hesson, Salt Lake City, OT 84114-6301, ee. Registrant must receive prior autho but is not limited to, change in location, it other state law, and specific provisions a fescribed in the application/amendment or implied, to registrant for issuance or d esued this 26th day of August, 2010, un ay	rization for any use or activity not species or numbers of animals. on complying with provisions in R tipulated herein. Activities author request submitted. This COR is t enial of future applications.	squire additional review by the Division will be subject to an amendment authorized under this COR or any rule pertaining thereto. This includes, 1657-3, Title 23, Utah Code, and all applicable foreign, federal, local, or fized harain must be carried out in accordance with and for the purposes valid only for the dates indicated herein and gives no rights, either express
Hesson, Sait Lake City, OF 84114-6301, lee. Registrant must receive prior autho but is not limited to, change in location, it The validity of this COR is dependent up other state law, and specific provisions a described in the application/amendment or implied, to registrant for issuance or d issued this 26th day of August, 2010, un ay	rization for any use or activity not opooles or numbers of animals. on complying with provisions in R tipulated herein. Activities author request submitted. This COR is t enial of future applications. der authority granted by R657-3 e	squire additional review by the Division will be subject to an amendment authorized under this COR or any rule pertaining thereto. This includes, 1657-3, Title 23, Utah Code, and all applicable foreign, federal, iocal, or ized harain must be carried out in accordance with and for the purposes valid only for the dates indicated herein and gives no rights, either express and Title 23, Utah Code.
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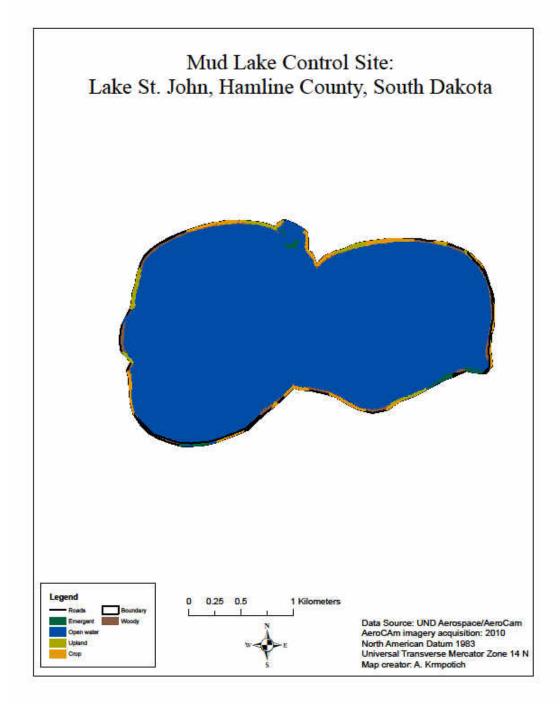
FISH 4 WILDLIFE UNIT	ED STATES DEPARTMENT OF THE FISH AND WILDLIFE SERVICE	INTERIOR	Lagarana concerción as 42	Credited Permit No. 51510 - 10-15	
Jan Ville	Benton Lake NWR		Date June 2	21, 2010	
	922 Bootlegger Trail Great Falls, MT 59404-6133		Period of Use (in	clusive)	
Contraction of the second	Phone (406)727-7400		From	June 21, 2010	
	SPECIAL USE PERM	MIT	То	July 31, 2010	
Permittee Name		Permittee Add	dress		
Annmarie Krmpot	ich	Universit	y of North Dal	kota	
Annmarie.Krmpot	ich@und.edu	Dept of E			
	14		orks, ND 5820	01	
3	*	(320)333	8-6678		
Purpose (specify in detail	privilege requested, or units of products in	ivolved)			
also segregated on the	e genetic variability among sub-population: wintering grounds and what landscape cha ated with breeding colonies.	aracteristics (co	over type, open wal	ter to emergent veg ratios	and
	numbers mater and bounds, or other reas	onizable desig	nations)		
Description (specify unit	numbers, metes and ocunos, or other reco				
장님은 아이는 아이는 것이 같아?		12/16/16/10/201612			
장님은 아이는 아이는 것이 같아?	in Units 5 on Benton Lake refuge will be sa	12/16/16/10/201612			
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Special Use Permit: Benton Lake National Wildlife Refuge, Montana

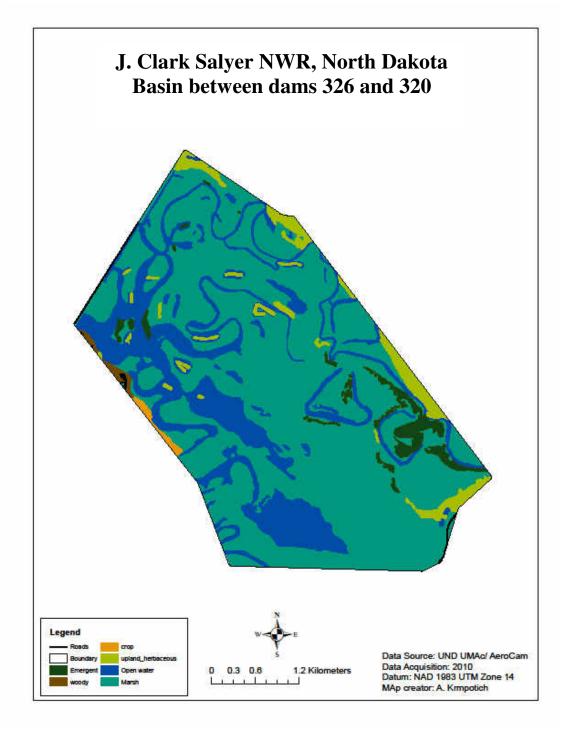
Appendix B Digitized Maps

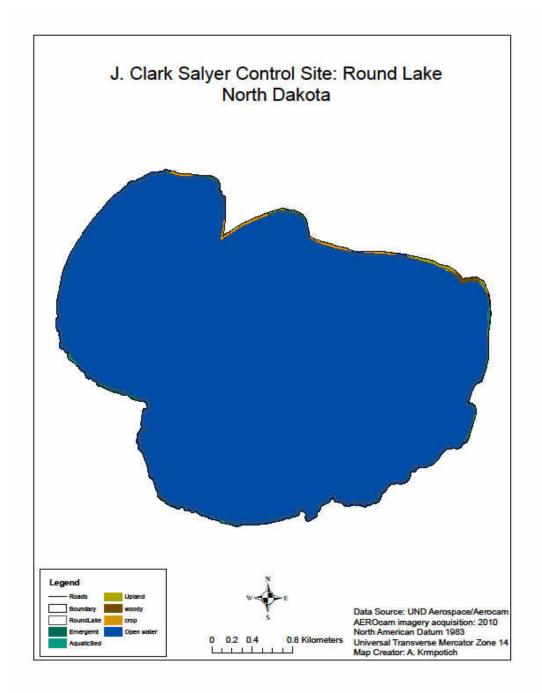
Mud Lake at Sand Lake National Wildlife Refuge, SD

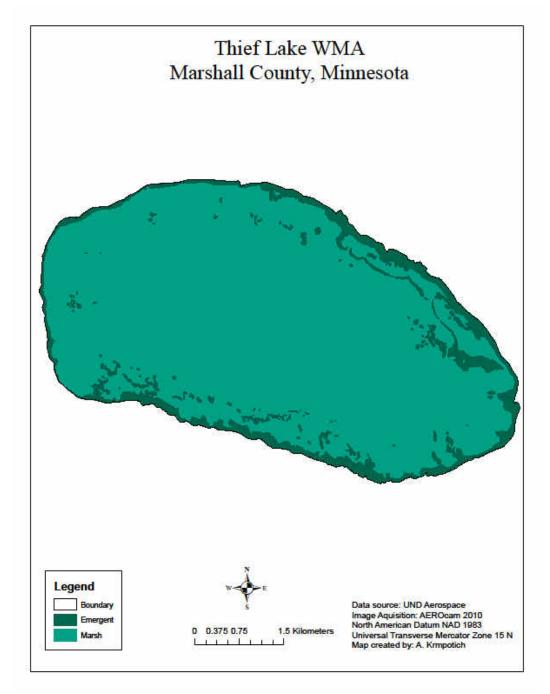


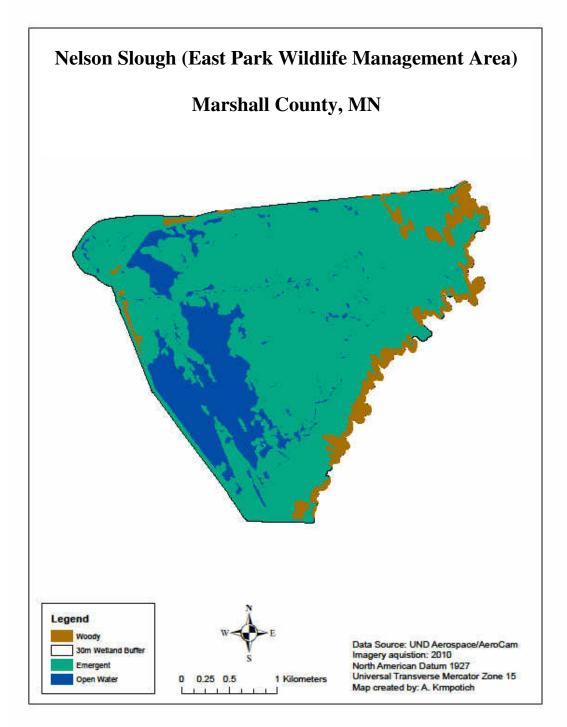


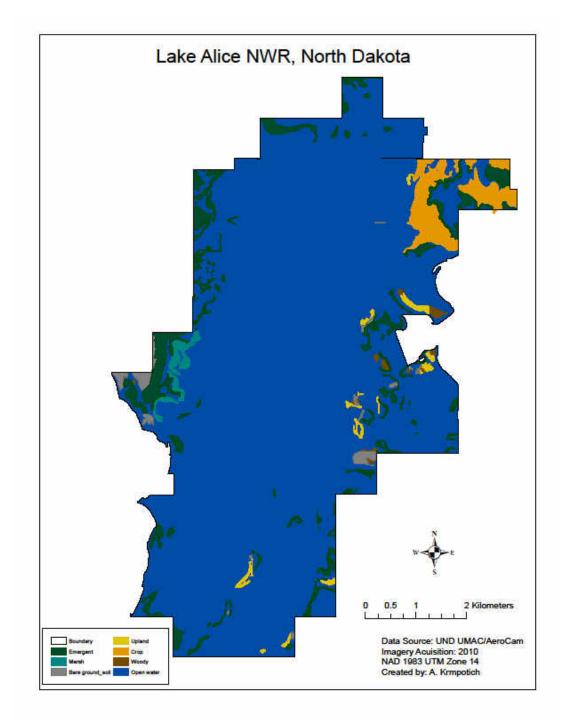
J. Clark Salyer National Wildlife Refuge, ND



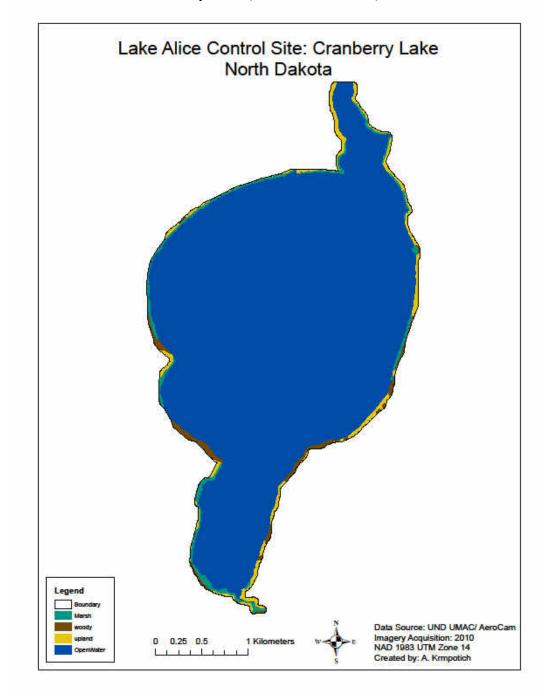








Lake Alice National Wildlife Refuge, ND



Cranberry Lake (Lake Alice Control), ND